

The role of urban green space design to support bird community in the urban ecosystem

DANA RISQY PRIHANDI, SANDY NURVIANTO*

Department of Forest Resource Conservation, Faculty of Forestry, Universitas Gadjah Mada. Jl. Agro No. 1, Bulaksumur, Sleman 55281, Yogyakarta, Indonesia. Tel.: +62-274-550541, *email: sandy_nurvi@yahoo.com

Manuscript received: 9 March 2022. Revision accepted: 26 March 2022.

Abstract. Prihandi DR, Nurvianto S. 2022. *The role of urban green space design to support bird community in the urban ecosystem. Biodiversitas 23: 2137-2145.* The increase in human population due to urbanization causes the expansion of urban areas that threatens forest lands to be lost and makes a serious threat to its biodiversity. Along with changes in the landscape due to urbanization, open green space is an alternative solution to maintaining biological diversity in urban areas. However, various types of open green spaces were developed following certain objectives of landscape management that potentially have different effects on biodiversity. This research aimed to identify the role of various open green spaces designs on bird biodiversity. Birds can be a good indicator of environmental quality because birds are a type of animal that can move in different habitat types and habitat areas and are often used as an indicator of habitat quality in urban areas. Bird observation was conducted using the point count method, while the environmental conditions were recorded using the nested and protocol sampling method. Those data were collected in five different open green spaces of Daerah Istimewa Yogyakarta Province. Species diversity shows the diversity of species in a bird community in a certain area using the Shanon-Wiener index. Bird communities were also analyzed using rarefaction analysis. This analysis was used to standardize the species abundance in each type of green open space to estimate the ideal species richness of open green space from the different number of samples. The Kruskal-Wallis test was used to compare the value of diversity, richness, and relative abundance of bird species among five different types of open green space. We used Canonical Correspondence Analysis (CCA) to visualize the bird's response toward environmental gradients. Generalized Linear Model (GLM) was employed to determine the effect of the environmental factors on the diversity of bird species. The result showed that different types of urban green spaces have different richness and relative abundance of bird species. We found 5925 individual birds from 40 species in five different types of open green spaces. The garden was the type of open green space with the highest bird species diversity. Of the many variables taken, the slope and the number of vehicles per minute affected the diversity of species. Those two variables were also negatively correlated with the diversity of bird species in the five types of open green spaces. The number of vehicles causes noise and pollution, which causes a decrease in bird populations. Flat slopes can also help birds get food, such as seeds on a flat surface. Creating more gardens for public and private spaces becomes an alternative solution to enhance bird diversity in the urban area.

Keywords: Bird diversity, garden, GLM, urbanization, urban ecology

INTRODUCTION

Urbanization is caused by the increasing human population inhabiting urban areas. The increase of the human population due to urbanization causes the expansion of urban areas that threatens forest lands to be lost and makes a serious threat to its biodiversity (Filloy et al. 2019). Forest destruction due to urbanization can threaten the existence of wild animals with species that are sensitive to their environment and wildlife habitats (Idilfitri and Mohamad 2012). Along with changes in the landscape due to urbanization, open green space is an alternative solution to maintaining biological diversity in urban areas.

Open green space is a space in urban areas whose cover is dominated by trees, shrubs, or grasses, either naturally or intentionally planted (Rakhshandehroo et al. 2017). The existence of open green space can increase the variety of landscape forms in urban areas with the presence of vegetation between cities. Open green space is an important component for biodiversity conservation in urban areas, as a sanctuary that allows certain species to survive

in urban areas (Leveau et al. 2019) and the important element of space for a city to maintain the quality of the city environment by controlling air pollution and water infiltration media (De La Barrera et al. 2016). As a supporter of urban ecosystems, open green space has various forms, such as urban parks, gardens, greenways, cemeteries, and agriculture (Carbó-Ramírez and Zuria 2011; Lestari et al. 2012).

Open green space in urban areas plays an important role in providing corridors to maintain the remaining native biodiversity (Callaghan et al. 2019; Fernández-juricic and Jokimäki 2001; Tryjanowski et al. 2017). However, the magnitude of the effect of open green space on biodiversity depends on the characteristics or design of open green space, such as management, cultivation, extent and form of natural vegetation, or even shrub and tree cover, which add to the complexity of habitats, nesting sites and food sources for birds (Korányi et al. 2020; MacGregor-Fors et al. 2010; Paker et al. 2014). This open green space allows the emergence of species that are less tolerant of disturbances

in urban landscapes (Carbó-Ramírez and Zuria 2011; Fernández-juricic and Jokimäki 2001).

There are various types of open green spaces that were developed following certain objectives of landscape management that potentially have different effects on biodiversity. This research aimed to identify the role of various open green spaces designs on bird biodiversity. Birds can be a good indicator of environmental quality because birds are a type of animal that can move in different habitat types and habitat areas and are often used as an indicator of habitat quality in urban areas. It is important to conduct research on bird species diversity in open green space to find out how good open green space is to protect biodiversity in urban areas.

MATERIALS AND METHODS

Study area

This research was conducted in open green spaces of the Yogyakarta Province, Indonesia. The research was

conducted on various types of open green spaces located in urban and suburban areas ($7^{\circ} 47' 49.4448''$ S, $110^{\circ} 22' 13.9044''$ E). The level of urbanization was determined by the percentage of a built-up area using the classification of Marzluff et al. (2001) with the output in the form of grids. The spatial data used for the delineation process in this study were Landsat 8 OLI/TIRS images with the help of QGIS Desktop 3.18 mapping software. This process began with processed Landsat 8 OLI/TIRS images, namely the atmospheric correction process or, in other words changing the reflectance at the satellite to surface reflectance. The next process was to quantify the percentage of the built-up area using the tools in the QGIS Desktop 3.18 software, namely the Semi-automatic Classification Plugin (SCP) and raster layer unique values report. After carrying out the process of quantifying the built-up area, the next process will determine the point on each open green space in the grid by looking at the land classification through images. We found 178 open green spaces that we used for observation points.

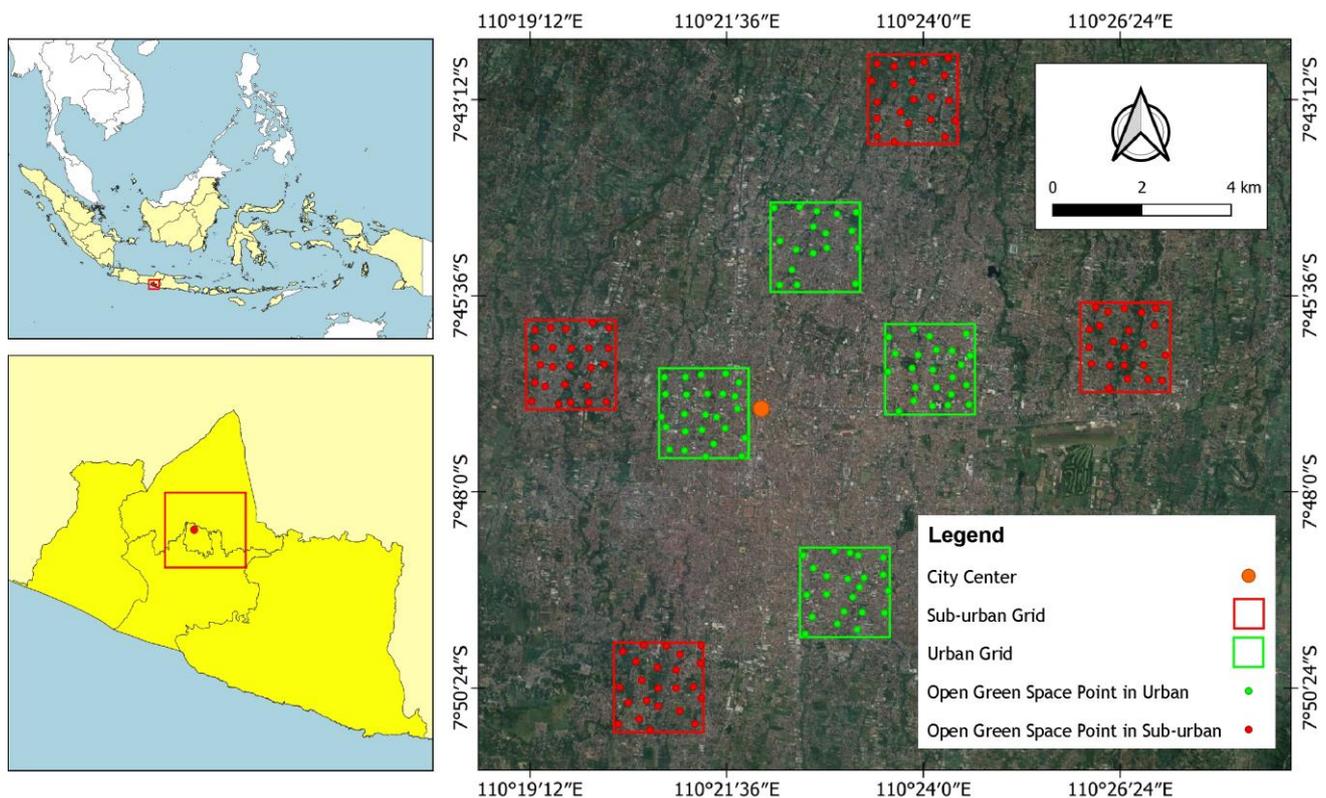


Figure 1. Location of Yogyakarta Province, Indonesia (*left*) and the sampling sites of this research that located in urban and sub-urban areas of Yogyakarta Province (*right*)

Table 1. Open green space area

Type of open green space	Number of points	Range area (ha)	Mean area (ha)
Garden	46	0.06-6.38	0.64
Agriculture	36	0.27-7.39	2.01
Greenway	32	0.06-1.05	0.47
Park	32	0.03-2.34	0.58
Cemetery	32	0.04-2.22	0.24

From 178 observation points observed, five different types of open green space were obtained, including gardens, agricultural land, parks, cemeteries, and greenways (Table 1). The type of open green space was identified by direct survey and observation on location. Garden was the most common open green space typically found in the study area with 46 points, followed by agriculture with 36 points and the parks, cemeteries, and greenways with 32 points. The area size of the five types of open green space was also different. The widest type of open green space is agriculture, with an average of 2.10 ha, while the open green space which has the smallest average area is the cemetery, which is 0.24 ha. Table 1 shows the range of each open green space area.

Procedures

Environmental data collection

We measured environmental data in each observation point. The variables that are collected are presented in Table 2, including biotic and abiotic variables. Vegetation data were collected using the nested sampling method. The nested sampling plot size used was 1 x 1 m for shrubs, 2 x 2 m for seedlings and undergrowth (height below 1.5 m), 5 x 5 m for saplings (diameter below 10 cm and height above 1.5 m), 10 x 10 m for poles (10-20 cm in diameter) and 20 x 20 m for trees (over 20 cm in diameter) (Oosting 1948). This method was developed to quantify and describe vegetation conditions. Variables that were collected are shown in Table 2. Protocol sampling was used to collect data on tree and shrub cover (Noon 1981). Slope, foliage density, distance from water, road, and city center were collected from the same points as the protocol sample. Human, vehicle, and predator presence was recorded by the used same protocol as bird observation method (Paker et al. 2014). Some research used these variables to identify environmental conditions in open green spaces (Lancaster and Rees 1979; MacGregor-Fors et al. 2010; Mörtberg and Wallentinus 2000; Mirski 2020; Paker et al. 2014; Shih 2018; Shwartz et al. 2008).

Bird data collection

Bird observations were carried out using the point count method. The points were determined by locating the observation points in each open green space's type with the distance between points being 400 m (Knutson et al. 2016). The grid design in this study is presented in Figure 1. Data collection was carried out in the morning at 06.00-10.00 WIB (GMT+7) and in the afternoon at 15.00-18.00 WIB (GMT+7) with the consideration that the peak of bird activity occurred at that time. The recording method was direct, and the observer used binoculars or a camera.

Table 2. List of environmental variables

Variable
• Number of tree species
• Number of shrub species
• Tree cover (%)
• Shrub cover (%)
• Foliage density (%)
• Height
▪ 0-30 cm
▪ 30-100 cm
▪ 100-200 cm
• Open Green Space's type
• Distance from water (m)
• Distance from road (m)
• Distance from city center (m)
• Human presence
• Vehicle presence
• Predator presence
• Slope (%)

Data analysis

The vegetation data that has been taken in each open green space is then processed to get the Important Value Index (IVI). IVI is a quantitative parameter used to determine the dominance of species in a plant community (Martono 2012). Bird species diversity was analyzed using the diversity index of Shannon-Wiener (Tramer 1969). Species diversity shows species diversity in a bird community in a certain area. The Kruskal-Wallis test was used to compare the value of bird diversity among different types of open green space (Mansor and Sah 2012). After knowing the results the significantly different from the Kruskal-Wallis test, Pairwise Wilcoxon-tests were used to find out which sites were different (Paker et al. 2014). The Canonical Correspondence Analysis (CCA) method was used to visualize bird responses to environmental factors (McGarigal et al. 2000; Morrison et al. 2001; Dray et al. 2003). The CCA method is very suitable for using data generated from complex data collection designs (McGarigal et al. 2000). Generalized Linear Model (GLM) analysis was employed to identify environmental factors affecting bird diversity in the open green spaces (Caprio et al. 2009). All statistical analyses were performed by R software version 4.1.1 with the Stats package.

RESULTS AND DISCUSSION

Vegetation structure and composition

In the five open green spaces, a total of 187 shrub species and 64 tree species were identified. The total of shrubs and trees species is shown in Figure 2. The shrubs that were commonly found in every open green space included *Acalypha indica*, *Ageratum conyzoides*, *Asystasia gangetica*, *Digitaria sanguinalis*, *Rivina humilis*, *Sphagneticola trilobata*, *Synedrella nodiflora*, and *Tridax procumbens*, while for tree species were *Artocarpus heterophyllus*, *Mangifera indica*, and *Muntingia calabura*.

In the agricultural type, we found 70 shrubs and 25 species of trees, with *D. sanguinalis* being dominated for shrub species and *Swietenia macrophylla* being dominated for trees species. In the garden type, there were 84 shrubs and 30 species of trees, with *A. gangetica* being the most common shrub species and *Falcataria moluccana* being the most common tree species. The greenways type found 47 species of shrubs and 30 species of trees. *Digitaria sanguinalis* was the most common shrub species, and *Pterocarpus indicus* was the most common tree species. We found 54 shrubs and 40 species in park type, with *Sida acuta* being the most common shrub species and *S. macrophylla* being the most common tree species. The last type of open green space, the cemetery type, contains 43 types of shrub species and 14 types of trees species.

Ageratum conyzoides was the common shrub species, and *M. indica* was the most common tree species.

Besides the important value index, tree cover and shrub cover could also show the structure and composition of vegetation in open green spaces. Of the five types of open green space, tree cover and shrub cover have similar values. In Figure 2, which shows the value of cover in the five open green spaces, the garden type has relatively higher shrub and tree cover than the other types. Almost all types of open green space have a higher tree cover value than shrub cover, except for the agricultural type (Figure 3). This fact is because agricultural open green spaces were used for agricultural functions for people and dominated by plants such as *Oryza sativa*.

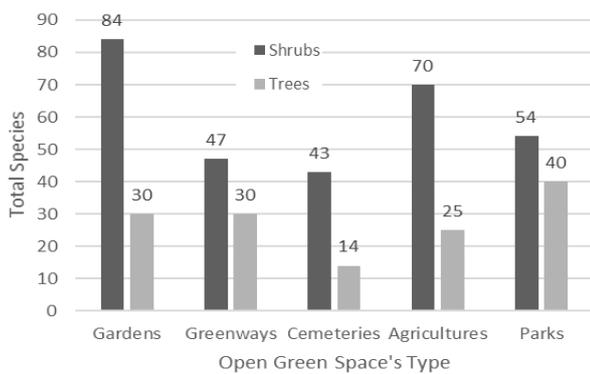


Figure 2. Total species of shrubs and trees in open green space

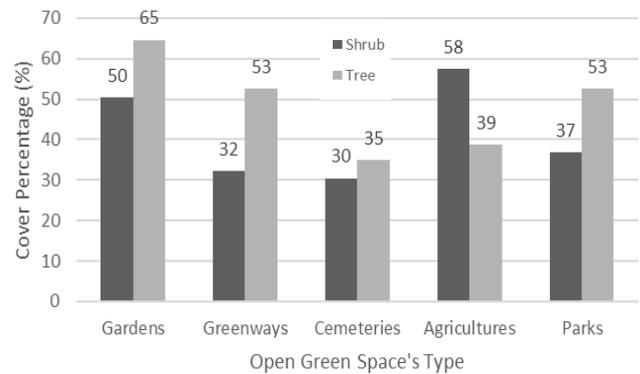


Figure 3. Shrub and tree cover in open green spaces (%)

Table 3. Shrub and tree species with Important Value Index (IVI)

Species name	Important Value Index (IVI)				
	Agricultural	Cemetery	Garden	Park	Greenway
Shrub species					
<i>Acalypha indica</i>	1.69	18.49	0.76	8.62	1.81
<i>Ageratum conyzoides</i>	6.08	20.98	7.43	10.08	8.52
<i>Asystasia gangetica</i>	11.82	13.43	21.64	1.54	4.90
<i>Chromoalena odorata</i>	12.86	4.17	9.65	-	7.44
<i>Desmodium triflorum</i>	3.37	-	3.15	9.15	6.71
<i>Digitaria sanguinalis</i>	23.72	12.86	4.08	11.01	27.07
<i>Rivina humilis</i>	3.16	2.73	3.54	6.32	11.43
<i>Sphagneticola trilobata</i>	9.73	9.29	5.77	7.98	12.53
<i>Synedrela nodiflora</i>	5.22	5.58	7.26	3.93	17.26
<i>Tridax procumbens</i>	3.55	7.97	2.31	1.17	11.26
Tree species					
<i>Artocarpus altilis</i>	32.32	21.67	23.99	-	2.90
<i>Artocarpus heterophyllus</i>	32.30	5.49	32.16	5.35	3.14
<i>Falcataria moluccana</i>	11.90	-	42.29	20.57	4.50
<i>Ficus benjamina</i>	-	5.47	-	27.04	40.04
<i>Mangifera indica</i>	23.50	63.63	23.18	15.60	5.05
<i>Plumeria sp.</i>	-	54.89	-	2.03	7.98
<i>Polyalthia longifolia</i>	19.70	-	1.74	6.85	30.00
<i>Pterocarpus indicus</i>	-	-	2.08	9.95	53.23
<i>Pterospermum javanicum</i>	13.39	8.28	26.54	-	17.41
<i>Samanea saman</i>	3.44	-	4.55	11.03	20.84
<i>Swietenia macrophylla</i>	44.19	-	35.88	31.62	28.23

Notes: The colored columns were the most common species in every Open Green Space's type. They were the five highest IVI in every Open Green Space's type

Bird species richness and diversity

The data collection results in five types of open green space found as many as 40 bird species with a total abundance of 5925 individual birds. Some species that are often found include Javan Munia (*Lonchura leucogastroides*), Eurasian Tree Sparrow (*Passer montanus*), and Yellow-vented Bulbul (*Pycnonotus goiavier*). Of the 40 species encountered, almost all species have a low conservation status (least concern/LC) according to the IUCN (BirdLife International 2016; BirdLife International 2016; BirdLife International 2017). The species that have a high conservation status according to the IUCN were the Bar-winged Prinia (*Prinia familiaris*) which was categorized as Near Threatened/NT (BirdLife International 2018), Javan Flameback (*Chrysocolaptes strictus*), which was categorized as Vulnerable/VU (BirdLife International 2016), and Javan myna (*Acridotheres javanicus*) which considered as Vulnerable/VU

(BirdLife International 2020). These three types are only found in one type of open green space with a few amounts. Of the five types of open green space, gardens had the highest species richness with 35 bird species, followed by agriculture (27), greenways (19), parks (18), and cemeteries (16). The agricultural type had the highest total abundance, namely 2169 individual birds, followed by the garden (1221), cemetery (982), park (788), and greenways (761) types, as can be seen in Table 4.

Species diversity showed the diversity or differences of species in a bird community in a certain area. The Shannon-Wiener diversity index (H') described bird diversity in each open green space. From the results obtained, the diversity of bird species in the five types of open green space was not significantly different. Garden was a type of open green space with the highest H' value of the other four types, as shown in Table 5.

Table 4. Bird species listed in Open Green Space's type

Species name	Conservation status	Open green spaces type				
		Agricultural	Garden	Park	Cemetery	Greenway
<i>Lonchura leucogastroides</i>	Least Concern/LC	+	+	+	+	+
<i>Lonchura punctulata</i>	Least Concern/LC	+	+	+	+	+
<i>Lonchura maja</i>	Least Concern/LC	+	+	+	+	+
<i>Passer montanus</i>	Least Concern/LC	+	+	+	+	+
<i>Orthotomus sepium</i>	Least Concern/LC	+	+	-	-	+
<i>Cisticola juncidis</i>	Least Concern/LC	+	+	-	+	+
<i>Todirhamphus chloris</i>	Least Concern/LC	+	+	+	+	+
<i>Halcyon cyanoventris</i>	Least Concern/LC	+	+	+	+	+
<i>Dicaeum trochileum</i>	Least Concern/LC	+	+	+	+	+
<i>Geopelia striata</i>	Least Concern/LC	+	+	+	+	+
<i>Hirundo rustica</i>	Least Concern/LC	+	-	-	-	+
<i>Ardeola speciosa</i>	Least Concern/LC	+	+	+	+	+
<i>Artamus leucorhynchus</i>	Least Concern/LC	+	+	-	-	+
<i>Pycnonotus aurigaster</i>	Least Concern/LC	+	+	+	+	+
<i>Prinia familiaris</i>	Near Threatened/NT	-	+	-	-	-
<i>Orthotomus sutorius</i>	Least Concern/LC	+	+	+	+	+
<i>Spilopelia chinensis</i>	Least Concern/LC	+	+	+	+	+
<i>Cacomantis merulinus</i>	Least Concern/LC	+	+	+	-	-
<i>Cacomantis sepulcralis</i>	Least Concern/LC	-	+	-	+	-
<i>Dendrocopos macei</i>	Least Concern/LC	-	+	-	-	-
<i>Aegithina tiphia</i>	Least Concern/LC	+	+	+	+	+
<i>Cinnyris jugularis</i>	Least Concern/LC	+	+	+	+	+
<i>Hirundo tahitica</i>	Least Concern/LC	+	-	+	+	-
<i>Prinia inornata</i>	Least Concern/LC	+	+	-	-	-
<i>Pycnonotus goiavier</i>	Least Concern/LC	+	+	+	+	+
<i>Zosterops palpebrosus</i>	Least Concern/LC	-	+	-	-	-
<i>Megalaima haemacephala</i>	Least Concern/LC	-	+	+	+	-
<i>Alcedo meninting</i>	Least Concern/LC	-	+	-	-	+
<i>Amaurornis phoenicurus</i>	Least Concern/LC	+	+	-	-	+
<i>Dendrocopos moluccensis</i>	Least Concern/LC	+	-	-	-	-
<i>Bubulcus ibis</i>	Least Concern/LC	+	+	-	-	-
<i>Treron vernans</i>	Least Concern/LC	-	+	+	-	-
<i>Chalcophaps indica</i>	Least Concern/LC	-	+	-	-	-
<i>Phaenicophaeus curvirostris</i>	Least Concern/LC	-	+	-	-	-
<i>Anthreptes malacensis</i>	Least Concern/LC	+	-	-	-	-
<i>Egretta garzetta</i>	Least Concern/LC	-	+	-	-	-
<i>Chrysocolaptes strictus</i>	Vulnerable/VU	-	+	-	-	-
<i>Aplonis minor</i>	Least Concern/LC	-	+	-	-	-
<i>Zapornia fusca</i>	Least Concern/LC	+	-	-	-	-
<i>Acridotheres javanicus</i>	Vulnerable/VU	-	-	+	-	-

Notes: +: Exist; -: Not exist

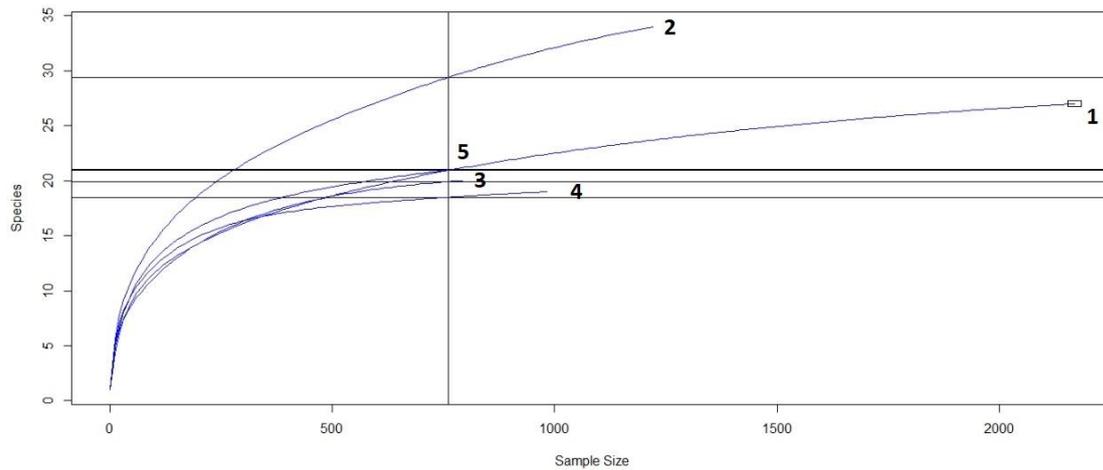


Figure 4. Rarefaction curve result. The curve showed the ideal species richness in every type of open green space. (Number 1: Agriculture land; Number 2: Garden, Number 3: Park, Number 4: Cemetery, Number 5: Greenway)

Table 5. Result of General Linear Model (GLM) Analysis

	Estimate	Std. Error	T value	Pr(> t)	
(Intercept)	1.3971767	0.0439090	31.820	<2e-16	***
Slope	-0.0099541	0.0045870	-2.170	0.0314	*
Vehicle/minute	-0.0019312	0.0009053	-2.133	0.0343	*
Null deviance		33.875			
Residual deviance		32.274			
AIC		209.2			

Note: P<0.05 symbolized by *, P<0.01 by **, P<0.001 by ***

Table 6. The value of species abundance, richness, and diversity in five types of Open Green Spaces

Open Green Space's type	Species abundance	Species richness	Species diversity (H')
Gardens	1221	34	2.1
Agricultures	2169	27	1.58
Parks	788	20	1.84
Cemeteries	982	19	1.91
Greenways	761	21	1.88

In theory, the difference in the number of plots could affect the species richness and abundance already counted. Therefore, bird communities were also analyzed using rarefaction analysis. In this analysis, standardization was used to standardize the species abundance in each type of green open space to estimate the ideal species richness of open green space (Irham 2015; James and Rathbun 1981; Kale 2014; Walker et al. 2008) so that the results were objective. The analysis was carried out by taking the type of green open space with the least number of plots as the standard, in this case, the greenway type. From the rarefaction curve results obtained, the highest bird species richness remained in the garden type with 29 species, followed by greenways (21), agriculture (20), gardens (19), and cemeteries (18). Gardens have a higher species richness than other types. This result follows the research conducted by Paker et al. (2014), which states that gardens can provide suitable habitats for many bird species. In

addition, species richness in a habitat provides food availability for bird communities (Tu et al. 2020).

Effect of environmental factors on bird community

The result of the Kruskal-Wallis analysis showed that the bird diversity among the types of open green spaces statistically was not significantly different (K=12,351, df=4, p-value=0.6778), while the bird species richness and abundance among the types of open green space were significantly different to species richness (K=12.351, df=4, p-value=0.01492) and bird species abundance (K=43.304, df=4, p-value<0.001). From the Pairwise Wilcoxon test, it was known that species richness in the types of agricultural open green spaces and greenways was significantly different, while the abundance of agricultural types with other open green spaces has a very significant difference.

CCA analysis was also carried out to visualize the response of the bird community to environmental factors related to the environmental aspect. It can be seen that along with the 1st (horizontal) axis, most bird species are located at the left-hand part of the axis (squares 1 and 4), where tree species, distance from water sources, and shrub cover indicate that these variables have a positive effect on bird community structure. Fewer bird species are located on the right-hand part of this axis (squares 2 and 3), where slope, presence of people, and predators are high, indicating that these variables have a negative effect on bird community structure.

distance from water sources, and shrub cover were positively correlated with bird communities in open green spaces (Figure 4). This result confirms the above statement about the number of tree species and high shrub cover formed a good bird community structure, while the distance from the water source showed that the farther the distance, the better the bird community in open green spaces. This event can happen because most bird species were found far from water sources or rivers. In addition, there were several variables such as the number of shrub species, presence of people, presence of predators, slopes, and canopy cover, which had a negative correlation in most bird communities but were positively correlated with some bird species. The presence of predators in question were animals that interfered with the presence of birds in the area, for example, cats and dogs, following the research of Boyle and Samson (1985) and Evans et al. (2009), which states that the presence of humans and predators has a negative correlation on the bird community. Humans usually visit public open green spaces such as parks with their pets in the form of cats or dogs (Paker et al. 2014). The study by Loss et al. (2013) even explained that cats in urban areas predate birds. The presence of people at the cemetery for grave pilgrimages can interfere with the presence of birds in the area. It was necessary to create an area in the open green space where humans and their pets did not have access to the place by planning the path in such a way that the relatively large area consists of shrubs that are not accessible to humans and their pets, and also with the rules for tying up pets (Paker et al. 2014).

From the regression analysis carried out, it was also known that the slope variable and the number of vehicles affected the diversity of bird species. Both had a negative correlation to the diversity of bird species. Urban areas traversed by vehicles can disturb bird communities. According to Kociolek et al. (2011), the number of vehicles causes noise and pollution, causing a reduction in bird populations. Kociolek et al. (2011) also said that passing vehicles could also result in bird deaths. Temporal adjustment of traffic flow and outreach to the public on increasing dependence on mass transit can be made to reduce road noise. In addition, the use of noise-absorbing tires can substantially reduce road noise levels. These features also increase people's quality of life and property value (Kociolek et al. 2011). Besides that, the slope also affected the diversity of bird species.

The slope had a negative correlation with the diversity of bird species, where the flatter the slope, the higher the diversity of bird species. This result was indicated because the research is in urban areas with flat slopes. In addition, open green spaces such as agriculture and gardens were mostly located on flat slope land (Sandström et al. 2006). In addition, flat slopes helped birds get food, such as grains on a flat surface.

Our results showed that open green spaces with high habitat quality, which in this study consisted of many species of shrub and trees, and shrub and tree cover also had an influence on bird communities. Therefore, when planning to build open green spaces, we recommend that designers avoid designs that do not have tree cover and do

not create monocultures of open green spaces without shrub cover. Open green space needs to be built with a composition consisting of various types of shrubs and trees, such as the garden in this study. The high number of shrub species and the number of tree species make the landscape structure more complex so that it will provide a wider spectrum of food sources for the bird community. In addition, open green space designers must consider things that can disturb bird communities, such as pollution and noise, by building vegetation walls around open green spaces to reduce pollution and noise. In addition, because of the negative effects of people and predators on bird community structures, we recommend creating shrub areas that will have limited access to people and their pets. Finally, there is a need for further research by combining urban planning experts with zoologists to continue to create new ideas to build cities with open green spaces that have good structure and habitat quality to conserve wildlife in urban areas.

ACKNOWLEDGEMENTS

We'd like to thank Ikhsan Fiqra Naufaliano, Nur Rochim, Abdiel Adam Utama, Eri Julian Catur Hidayat, and M. Akmal Ramadhan who help all the thing that we needed for the observation. Thanks to Fauzi Hamdan that already manage our fund. We also like to thank all the people of Daerah Istimewa Yogyakarta Province that really supportive when we took the data at their place. Thank you to Universitas Gadjah Mada for supporting this research. This research was funded by *Program Rekognisi Tugas Akhir*, Universitas Gadjah Mada, Yogyakarta, Indonesia (Universitas Gadjah Mada Final Assignment Recognition Program).

REFERENCES

- BirdLife International. 2016. *Chrysocolaptes strictus*. The IUCN Red List of Threatened Species 2016: e.T22726560A94925352. DOI: 10.2305/IUCN.UK.2016-3.RLTS.T22726560A94925352.en.
- BirdLife International. 2016. *Lonchura leucogastroides*. The IUCN Red List of Threatened Species 2016: e.T22719809A94645559. DOI: 10.2305/IUCN.UK.2016-3.RLTS.T22719809A94645559.en.
- BirdLife International. 2016. *Pycnonotus goiavier*. The IUCN Red List of Threatened Species 2016: e.T22712731A94346183. DOI: 10.2305/IUCN.UK.2016-3.RLTS.T22712731A94346183.en.
- BirdLife International. 2017. *Passer montanus* (amended version of 2017 assessment). The IUCN Red List of Threatened Species 2017: e.T22718270A119216586. DOI: 10.2305/IUCN.UK.2017-3.RLTS.T22718270A119216586.en.
- BirdLife International. 2018. *Prinia familiaris*. The IUCN Red List of Threatened Species 2018: e.T22713598A131851023. DOI: 10.2305/IUCN.UK.2018-2.RLTS.T22713598A131851023.en.
- BirdLife International. 2020. *Acridotheres javanicus*. The IUCN Red List of Threatened Species 2020: e.T103871334A176499647. DOI: 10.2305/IUCN.UK.2020-3.RLTS.T103871334A176499647.en.
- Boyle SA, Samson FB. 1985. Effects of nonconsumptive recreation on wildlife: a review. *Wildl Soc Bull* 13 (2): 110-116. DOI: 10.2307/3781422.
- Callaghan CT, Bino G, Major RE, Martin JM, Lyons MB, Kingsford RT. 2019. Heterogeneous urban green areas are bird diversity hotspots: insights using continental-scale citizen science data. *Landsc Ecol* 34 (6): 1231-1246. DOI: 10.1007/s10980-019-00851-6.

- Callaghan CT, Major RE, Lyons MB, Martin JM, Kingsford RT. 2018. The effects of local and landscape habitat attributes on bird diversity in urban greenspaces. *Ecosphere* 9 (7). DOI: 10.1002/ecs2.2347.
- Caprio E, Ellena I, Rolando A. 2009. Assessing habitat/landscape predictors of bird diversity in managed deciduous forests: A seasonal and guild-based approach. *Biodivers Conserv* 18 (5): 1287-1303. DOI: 10.1007/s10531-008-9478-1.
- Carbó-Ramírez P, Zuria I. 2011. The value of small urban greenspaces for birds in a Mexican city. *Landsc Urban Plann* 100 (3): 213-222. DOI: 10.1016/j.landurbplan.2010.12.008.
- Chamberlain DE, Gough S, Vaughan H, Vickery JA, Appleton GF. 2007. Determinants of bird species richness in public green spaces. *Bird Stud* 54 (1): 87-97. DOI: 10.1080/00063650709461460.
- Daniels GD, Kirkpatrick JB. 2006. Does variation in garden characteristics influence the conservation of birds in suburbia? *Biol Conserv* 133 (3): 326-335. DOI: 10.1016/j.biocon.2006.06.011.
- De La Barrera F, Reyes-Paecke S, Banzhaf E. 2016. Indicators for green spaces in contrasting urban settings. *Ecol Indic* 62, 212-219. DOI: 10.1016/j.ecolind.2015.10.027.
- Dray S, Chessel D, Thioulouse J. 2003. Co-inertia analysis and the linking of ecological data tables. *Ecology* 84 (11): 3078-3089. DOI: 10.1890/03-0178.
- Evans KL, Newson SE, Gaston KJ. 2009. Habitat influences on urban avian assemblages. *Ibis* 151 (1): 19-39. DOI: 10.1111/j.1474-919X.2008.00898.x.
- Fernández-juricic E, Jokimäki J. 2001. A habitat island approach to conserving birds in urban landscape: Case studies from southern and northern Europe. *Biodivers Conserv* 10 (12): 2023-2043.
- Fillooy J, Zurita GA, Bellocq MI. 2019. Bird diversity in urban ecosystems: the role of the biome and land use along urbanization gradients. *Ecosystems* 22 (1): 213-227. DOI: 10.1007/s10021-018-0264-y.
- Idilfitri S, Mohamad NHN. 2012. Role of ornamental vegetation for birds' habitats in urban parks: Case study FRIM, Malaysia. *Procedia - Soc Behav Sci* 68, 894-909. DOI: 10.1016/j.sbspro.2012.12.275.
- Irhani M. 2015. Komunitas burung bawah tajuk di hutan perbatasan, Kabupaten Nunukan, Kalimantan Utara. *Zoo Indonesia* 24 (1): 1-14. https://e-journal.biologi.lipi.go.id/index.php/zoo_indonesia/article/view/737
- James FC, Rathbun S. 1981. Rarefaction, relative abundance, and diversity of avian communities. *The Auk* 98 (5): 785-800. <http://www.jstor.org/stable/10.2307/4085899> %5Cnhttp://library.unm.edu/sora/Auk/v098n04/p0785-p0800.pdf%5Cnhttp://www.jstor.org/stable/10.2307/4085899
- Kale MA. 2014. Bird Species in Urban and Bird diversity patterns along an urbanisation gradient and crop damage caused by birds on the Deccan Plateau, India (Issue November).
- Knutson MG, O'Brien L, Sutherland TW, Carlyle KL, Herner-Thogmartin J, Carter L. 2016. National protocol framework for the inventory and monitoring of breeding landbirds using point counts, Version 2.0. February, 81.
- Kociolek AV, Clevenger AP, St. Clair CC, Proppe DS. 2011. Efectos de las Redes de Caminos sobre Poblaciones de Aves. *Conserv Biol* 25 (2): 241-249. DOI: 10.1111/j.1523-1739.2010.01635.x.
- Korányi D, Gallé R, Donkó B, Chamberlain DE, Batáry P. 2021. Urbanization does not affect green space bird species richness in a mid-sized city. *Urban Ecosyst* 24 (4): 789-800. DOI: 10.1007/s11252-020-01083-2.
- Lancaster RK, Rees WE. 1979. Bird communities and the structure of urban habitats. *Can J Zool* 57 (12): 2358-2368. DOI: 10.1139/z79-307.
- Lestari SP, Noor I, Ribawanto H. 2012. Dalam upaya mewujudkan sustainable city (Studi pada masterplan pengembangan RTH Tahun 2012-2032 di Kabupaten Nganjuk). *Jurnal Administrasi Publik (JAP)* 2 (3): 381-387.
- Leveau LM, Ruggiero A, Matthews TJ, Isabel Bellocq M. 2019. A global consistent positive effect of urban green area size on bird richness. *Avian Res* 10 (1): 1-14. DOI: 10.1186/s40657-019-0168-3.
- Loss SR, Will T, Marra PP. 2013. The impact of free-ranging domestic cats on wildlife of the United States. *Nat Commun* 4 (1): 1-8. DOI: 10.1038/ncomms2380.
- MacGregor-Fors I, Morales-Pérez L, Schondube JE. 2010. Migrating to the city: Responses of neotropical migrant bird communities to urbanization. *Condor* 112 (4): 711-717. DOI: 10.1525/cond.2010.100062.
- Mansor MS, Sah SAM. 2012. The influence of habitat structure on bird species composition in lowland Malaysian rain forests. *Trop Life Sci Res* 23 (1): 1-14.
- Martono DS. 2012. Analisis vegetasi dan asosiasi antara jenis-jenis pohon utama penyusun hutan tropis dataran rendah di Taman Nasional Gunung Rinjani Nusa Tenggara Barat. *Jurnal Agri-Tek* 13 (2): 18-27.
- Marzluff JM, Bowman R, Donnelly R. 2001. A historical perspective on urban bird research: trends, terms, and approaches. *Avian Ecol Conserva Urbanizing World* 1-17. DOI: 10.1007/978-1-4615-1531-9_1.
- Mason J, Moorman C, Hess G, Sinclair K. 2007. Designing suburban greenways to provide habitat for forest-breeding birds. *Landsc Urban Plann* 80 (1-2): 153-164. DOI: 10.1016/j.landurbplan.2006.07.002.
- Mayorga I, Bichier P, Philpott SM. 2020. Local and landscape drivers of bird abundance, species richness, and trait composition in urban agroecosystems. *Urban Ecosyst* 23 (3): 495-505. DOI: 10.1007/s11252-020-00934-2.
- McGarigal K, Stafford S, Cushman S. 2000. Canonical Correlation Analysis. *Multivariate Statistics for Wildlife and Ecology Research*. Springer, New York.
- Mirski P. 2020. Tree cover density attracts rare bird of prey specialist to nest in urban forest. *Urban For Urban Green* 55, 1-4. DOI: 10.1016/j.ufug.2020.126836.
- Morrison ML, Block WM, Strickland MD, Collier BA, Peterson MJ. 2001. *Wildlife study design*. Springer Verlag New York, Inc.
- Mörtberg U, Wallentinus HG. 2000. Red-listed forest bird species in an urban environment - Assessment of green space corridors. *Landsc Urban Plann* 50 (4): 215-226. DOI: 10.1016/S0169-2046(00)00090-6.
- Noon BR. 1981. Techniques for sampling avian habitats. The use of multivariate statistics in studies of wildlife habitat (DE Capen ed.). *USDA For Serv Gen Tech Rep RM-87*: 42-52.
- Oosting HJ. 1948. *The Study of Plant Communities*. W. H. Freeman and Company, San Francisco.
- Ortega-Álvarez R, MacGregor-Fors I. 2009. Living in the big city: Effects of urban land-use on bird community structure, diversity, and composition. *Landsc Urban Plann* 90 (3-4): 189-195. DOI: 10.1016/j.landurbplan.2008.11.003.
- Paker Y, Yom-Tov Y, Alon-Mozes T, Barnea A. 2014. The effect of plant richness and urban garden structure on bird species richness, diversity and community structure. *Landscape Urban Plann* 122, 186-195. DOI: 10.1016/j.landurbplan.2013.10.005.
- Pirzio Biroli A, Van Doren BM, Grabowska-Zhang A. 2020. Drivers of avian species richness and community structure in urban courtyard gardens. *J Urban Ecol* 6 (1): 1-11. DOI: 10.1093/jue/juz026.
- Rakhshandehroo M, Afshin S, Mohd Yusof MJ. 2017. Terminology of urban open and green spaces. 11th ASEAN Postgraduate Seminar, Malaysia. APGS 2017, November.
- Sandström UG, Angelstam P, Mikusiński G. 2006. Ecological diversity of birds in relation to the structure of urban green space. *Landsc Urban Plann* 77 (1-2): 39-53. DOI: 10.1016/j.landurbplan.2005.01.004.
- Shah SB, Sharma HP. 2022. Bird diversity and factors affecting bird abundance at Dullu Municipality, Dailekh, Nepal. *Biodiversitas* 23 (3): 1535-1545. DOI: 10.13057/biodiv/d230343.
- Shih WY. 2018. Bird diversity of greenspaces in the densely developed city centre of Taipei. *Urban Ecosyst* 21 (2): 379-393. DOI: 10.1007/s11252-017-0720-z.
- Shwartz A, Shirley S, Kark S. 2008. How do habitat variability and management regime shape the spatial heterogeneity of birds within a large Mediterranean urban park? *Landsc Urban Plann* 84 (3-4): 219-229. DOI: 10.1016/j.landurbplan.2007.08.003.
- Tramer EJ. 1969. Bird species diversity: Components of Shannon's formula. *Ecology* 50 (5): 927-929. DOI: 10.2307/1933715.
- Tryjanowski P, Morelli F, Mikula P, Krištín A, Indykiewicz P, Grzywaczewski G, Kronenberg J, Jerzak L. 2017. Bird diversity in urban green space: A large-scale analysis of differences between parks and cemeteries in Central Europe. *Urban For Urban Green* 27, 264-271. DOI: 10.1016/j.ufug.2017.08.014.
- Tu HM, Fan MW, Ko JJC. 2020. Different habitat types affect bird richness and evenness. *Sci Rep* 10 (1): 1-10. DOI: 10.1038/s41598-020-58202-4.
- Walker SC, Poos MS, Jackson DA. 2008. Functional rarefaction: Estimating functional diversity from field data. *Oikos* 117 (2): 286-296. DOI: 10.1111/j.2007.0030-1299.16171.x.
- Zhang Z, Huang G. 2020. How do urban parks provide bird habitats and birdwatching service? Evidence from Beijing, China. *Remote Sens* 12 (19): 1-15. DOI: 10.3390/rs12193166.