

Predicting of Komodo dragon's potential prey habitat suitability using MaxEnt in Riung Nature Reserve, Flores, East Nusa Tenggara, Indonesia

FADLAN PRAMATANA^{1,2,*}, YUSRATUL AINI¹, NIXON RAMMANG^{1,2}, YOSEP SERAN MAU^{2,3},
I G.B. ADWITA ARSA^{2,3}, ARIEF MAHMUD⁴

¹Department of Forestry, Faculty of Agriculture, Universitas Nusa Cendana. Jl. Adisucipto, Penfui, Kupang 85001, East Nusa Tenggara, Indonesia.
Tel./Fax.: +62-380-881085, *email: fadlan.pramatana@staf.undana.ac.id

²Archipelagic Drylands Laboratory, Universitas Nusa Cendana. Jl. Adisucipto, Penfui, Kupang 85001, East Nusa Tenggara, Indonesia

³Department of Agrotechnology, Faculty of Agriculture, Universitas Nusa Cendana. Jl. Adisucipto, Penfui, Kupang 85001, East Nusa Tenggara, Indonesia

⁴Natural Resources Conservation Agency of East Nusa Tenggara Province. Jl. SK Lerik Kelapa Lima, Kupang 85001, East Nusa Tenggara, Indonesia

Manuscript received: 12 May 2023. Revision accepted: 7 June 2023.

Abstract. Pramatana F, Aini Y, Rammang N, Mau YS, Arsa IGBA, Mahmud A. 2023. Predicting of Komodo dragon's potential prey habitat suitability using MaxEnt in Riung Nature Reserve, Flores East Nusa Tenggara. *Biodiversitas* 24: 3128-3139. The Komodo dragon (*Varanus komodoensis* Ouwens, 1912) is a big lizard species from the Varanidae family that belongs to the Endangered category (EN) listed on the IUCN red list and Appendix I CITES. This study aimed to reveal the distribution of potential preys Komodo dragons in Rinca Island, Komodo National Park, Manggarai District, East Nusa Tenggara, Indonesia, using Maximum Entropy (MaxEnt), which was collected using rapid assessment methods. The presences of the Komodo dragon's potential prey come from direct and indirect observation or previous studies. We collected 510 points of Komodo dragon prey presence in Riung, Ngada District, East Nusa Tenggara, Indonesia from six species, including cattle, but only used 127 points for analysis based on the correlation. Long-tailed macaque, wild boar, civet, Timor deer, feral horses, and cows were the potential prey for komodo in Riung. Most of the points come from cattle, such as cows. On the other hand, we used environmental habitat to represent prey habitats such as elevation, slope, land surface temperature, moisture index, vegetation index, and distance from specific objects such as distance from agriculture, rivers, road, savanna, and settlement. Komodo dragon's potential prey in Riung was distributed in savanna, mangrove, and lowland forest. The result showed three suitable habitats for the Komodo dragon's potential prey dominated by low and moderate-suitability areas.

Keywords: Habitat suitability, komodo dragon, maxent, prey, Riung

INTRODUCTION

Species distribution models estimate the relationship between environmental characteristics at species occurrences and environmental characteristics within the species' general area of occurrence (Franklin 2010). Hence, predictive models of potential geographic distributions are widely used for various applications in ecology, conservation, and biogeography (Graham et al. 2004; Guisan and Thuiller 2005). MaxEnt is one of the algorithms that can be used to predict a species' potential distribution (Phillips 2005). It is a machine-learning approach based on presence-only data, evaluating the likelihood of presence in a given cell based on environmental features in the same cell (Elith et al. 2006; Wisz et al. 2008; Elith et al. 2010; Elith et al. 2011).

Climate change is projected to match these drivers in intensity and possibly outpace them in the next 50 years (Newbold 2018). Future climate change can lead to shifts in the distribution and abundance of species (Thomas et al. 2004; Ehrlén and Morris 2015; Wang et al. 2018), extinction of species populations (Thomas et al. 2004; Keith et al. 2008; Bestion et al. 2015), range shifts (Chen et al. 2011; Nenzén and Araújo 2011; Bellard et al. 2012;

Iverson and McKenzie 2013; Habibzadeh et al. 2021) phonological changes (Anenkhonov 2009; Cuena-Lombrana et al. 2018; Merilä and Hendry 2014; Wolkovich et al. 2012), and physiological trait changes (Dillon et al. 2010; Fois et al. 2018).

Studying the effects of future climate change on the distribution of species is one of the fundamentals of managing informative activities for the conservation of biodiversity (Hosseinzadeh et al. 2017; Kaky and Gilbert 2017, 2019; Chefaoui et al. 2018; Hosseinzadeh et al. 2018, 2020; Baker et al. 2021; Fathinia et al. 2020), and studying the genus *Varanus* (Malakhov and Chirikova 2018; Shadloo et al. 2021; Baral et al. 2023).

The Komodo dragon (*Varanus komodoensis* Ouwens, 1912) is the largest living lizard from the Varanidae family that belongs to the Endangered category (EN), which is listed on the IUCN red list (Jessop et al. 2021). In addition, the Komodo dragon's local name is *ora*, a flagship species in the Komodo National Park area listed in Appendix I of CITES, indicating that selling this species is prohibited in a state of life and body parts. Komodo dragons are only spread across four populations within the Komodo National Park area, with the largest populations on Komodo Island, Rinca Island, Gili Motang, and Nusa Kode, East Nusa

Tenggara, Indonesia. (Ciofi and De Boer 2004; Jessop et al. 2007; Purwandana et al. 2014; Ariefiandy et al. 2017). Santosa et al. (2012) reported the distribution of dragons on Rinca Island often clustered in certain places, such as guard posts or forest areas during the dry season.

In 1980 the Komodo National Park was established to protect most of the Komodo dragon distribution, including the islands of Komodo, Rinca, Padar, Gili Motang, and Nusa Kode. Furthermore, in 1985 several Nature Reserve Areas (*Kawasan Suaka Alam/KSA*) were established to protect the distribution area of the Komodo dragon on Flores Island: Wae Wuul Nature Reserve (*Cagar Alam/CA*) on the west coast, Wolo Tado CA, Riung CA and Seventeen Islands TWA on the north coast. These four locations are under the Natural Resources Conservation Agency (*Balai Besar Konservasi Sumber Daya Alam/BBKSDA*) management of East Nusa Tenggara Province (Nusa Tenggara Timur/NTT). Many habitat pockets are scattered on the island of Flores, East Nusa Tenggara, Indonesia, only classified as having a low population density (Ciofi and De Boer 2004).

Threats to the existence of Komodo dragons include habitat degradation (Ariefiandy et al. 2021) and potentially invasive animals, such as Black-Spined Toad (*Duttaphrynus melanostictus* Schneider, 1799), which are feared to enter the Komodo dragon's habitat (Kennedi et al. 2020; Ujvari et al. 2015). In addition, hunting for Komodo dragons, baiting (Santosa et al. 2012), and massive development also threaten their existence. Ardiantiono et al. (2018) reported that there are long-term consequences when the natural presence of Komodo dragons comes into contact with tourism activities. Komodo dragons probably will lose their instinct to hunt and start consistently expecting leftover food or food visitors provide in tourist areas. Walpole (2001) reported that additional feeding was carried out to attract tourists to see Komodo dragons at viewing sites. Komodo dragons show habituation by reducing their negative responses to humans (Ardiantiono et al. 2018; Fauzia 2020). In addition, Ardiantiono (2018)

also explained that providing additional food would have a long-term effect on the Komodo dragons.

Mustari et al. (2011) state that five large mammals are prey animals for the Komodo dragon: Timor deer, water buffalo, wild horses, wild boars, and long-tailed monkeys found on Rinca Island. Meanwhile, since 2008, the NTT BBKSDA, with the Komodo Survival Program (KSP), has studied the population of Komodo dragons and their prey animals in the Wae Wuul CA area. They assumed that when an animal is found as prey for the Komodo dragon, this indicates the presence of the Komodo dragon. However, the distribution of prey animals on Rinca Island is random, with a quite large number (Mustari et al. 2011). Furthermore, Timor deer, water buffalo, and horses are herbivores, so the presence of forage as feed plays a role in the existence and sustainability of these animal populations. Therefore, this research was proposed to model the presence of prey animals in the NTT BBKSDA area, which is considered a meeting point for the Komodo dragons. This research also supports the conservation of priority animals currently threatened by land use change so that the results of spatial modeling of habitat suitability as a consideration for managers to manage areas with high habitat suitability for Komodo dragon's potential prey. In addition, the results of this study also provide an overview of the suspected existence of Komodo dragons whose habitat continues under pressure.

MATERIALS AND METHODS

Study area

This study was conducted from July to December 2022 in Riung Nature Reserve in Ngada Regency, East Nusa Tenggara, Indonesia from 08.00-17.00 WITA. Those areas are within the management of the East Nusa Tenggara Provincial Natural Resources Conservation Agency (*Balai Besar Konservasi Sumber Daya Alam/BBKSDA*) (Figure 1).

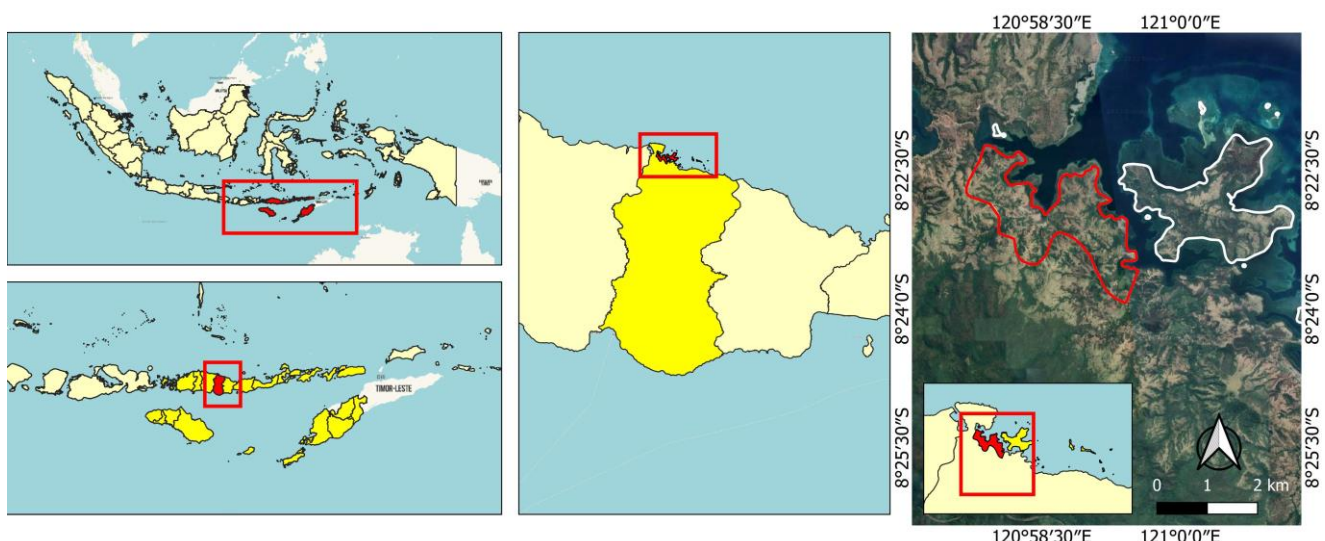


Figure 1. Location of Komodo dragon's potential prey in Riung Nature Reserve Ngada District, East Nusa Tenggara, Indonesia

Procedures

Data collection

The research was started by conducting a literature study on the distribution of Komodo dragons, their prey animals, and their habitats to facilitate making tally sheets. Then we collected the presence of the Komodo dragon's potential prey from direct and indirect observation or previous studies. Finally, observations were made using rapid assessment methods in Riung Nature Reserve. Rapid assessment is an observation by recording species found directly or indirectly, such as mammal footprints and feces, without any special paths or locations (Bismark 2011). In addition, observers record the mammal types found during site surveys or walking outside observation hours (Bismark 2011).

Furthermore, the coordinate points are taken at the direct or indirect object's present location as primary data in the subsequent spatial analysis (Merow et al. 2013; Nezer et al. 2017; Phillips 2017; Phillips et al. 2017). Coordinate point data collection using tally sheets compiled on KoboCollect, a web-based and smart mobile app being easier to access, less expensive, and more efficient for data collection in many scientific fields (Lakshminarasimhappa 2022). For example, collect is used in the socio-economic and socio-cultural (Deniau et al. 2017), used for controlling animal problems (Le Bel et al. 2016), used for wildlife management (Palla et al. 2016), etc.

Coordinate points of the observed result were then extrapolated to produce a map of habitat suitability index with ArcGIS 10.8.2 (licensed to Fadlan Pramatana) to create environmental variables and Maxent to create a spatial model (Çoban et al. 2020; Merow et al. 2013; Phillips et al. 2017). In addition, data on the presence of prey animals for the Komodo dragon were also obtained from literature studies carried out by individuals or institutions. The point of presence of the Komodo dragon's potential prey will be collected in a .csv file in Microsoft Excel 2021 version for further analysis on Maxent (Phillips 2017).

Environmental variables

Environmental variables are needed to conduct the analysis to develop a species distribution model in Maxent (Phillips 2017; Phillips et al. 2006, 2017; Phillips and Dudík 2008). The environmental variables used in this

study include elevation, slope, land surface temperature, moisture index, vegetation index, and distance from specific objects, such as distance from agriculture, river, road, savanna, and settlement. The type of data, source of data, and data collection method are shown in Table 1.

Environmental variables would be used to predict the suitability habitat of the Komodo dragon's potential prey, representing ecological conditions (Williams et al. 2012). The selection of environmental variables is important to improve the accuracy of the potential distribution of species (Li and Ding 2016). Environmental variables can also be based on habitats or variables that affect the presence of these species (Remya et al. 2015; Ma and Sun 2018; Sharma et al. 2018; Zhang et al. 2019b; Li et al. 2023), etc.

Data analysis

Analyzing data for habitat suitability modeling uses two data types: the presence data for the Komodo dragon's potential prey and environmental variables (Phillips et al. 2006, 2017; Phillips and Dudík 2008; Phillips 2017). Data on the whereabouts of the Komodo dragon's potential prey are stored in comma-separated value (CSV) format, and continuous data for environment variables in ASCII (asc) format to run into maxent (Phillips 2017). Maxent is a tool to model species distribution with habitat suitability outcomes supported by environmental variables data (Che et al. 2014; Na et al. 2018; Sharma et al. 2018; Zhang et al. 2019b; Sun et al. 2021; Ab Lah et al. 2021). Before the maxent analysis is carried out, first to perform spatial autocorrelation testing at the present point and environmental variables to eliminate multicollinearity, which aims to improve the accuracy of the potential distribution model Songchitruksa and Zeng 2010; Merow et al. 2013; Brown 2014; Yu et al. 2016; Li and Ding 2016; Brown et al. 2017; Oxoli et al. 2017; Baek et al. 2019; Perkins-Taylor and Frey 2020). Shrestha (2020) stated that multicollinearity causes significant variables to become insignificant in statistical analysis. Multicollinearity analysis was performed in Microsoft Excel with a value less than -0.5 (negative), and more than 0.5 (positive), such as the analysis conducted by McCarthy (2015), which previously had 14 candidate environmental variables, after completing a multicollinearity analysis, in the final environmental variables used for Maxent analysis were only six.

Table 1. Data sources and collection methods for environmental variables of this study

Data source	Collection method	Environmental variables data
Digital Elevation Model	Download from DEMNAS	Elevation
Landsat 8	Download from Earthexplorer.usgs.gov	Slope
		Land surface temperature
		Moisture index
		Vegetation index
Ministry of Environment and Forestry webgis	Download from webgis	Savanna land cover
Indonesian topographical basemap	Download from inageoportal	Agriculture land cover
		River
		Road
Open street map data	Download from Openstreetmap.org	Settlement

After passing the multicollinearity test stages, all point data and environment variables are entered into the MaxEnt application. The settings in the Maxent application use the display by default. The MaxEnt application will provide 30% of the data as testing data and 70% as training data (Yang et al. 2013; Rahmati et al. 2016; Kornejady et al. 2017; Cabrera and Lee 2020). For maxent analysis, this study used ten times replication, and some studies also use ten times replication to get the best result (Khanum et al. 2013; Su et al. 2021; Zhang et al. 2019a) although the replication is an option that can be used to do multiple runs (Phillips 2017). In addition, the replication can display a statistical summary from spatial distribution modeling (avg, SD, etc.) (Phillips 2017).

RESULTS AND DISCUSSION

Spatial autocorrelation and multicollinearity test

During the study, we found 510 data from six species of Komodo dragon's potential prey in Riung, which were dominated by indirect object recordings in the form of feces such as house gecko, civet, Timor deer, horse, cow, and long-tailed macaque (Figure 2).

Furthermore, a hot spot analysis (Getis-Ord G_i^*) was carried out using ArcMap software to reduce spatial autocorrelation from object points data obtained (Oxoli et al. 2017; Songchitruksa and Zeng 2010). The first introduced Getis-Ord G_i^* is a distance statistic identifying spatial data patterns classified into hot and cold spots (Getis and Ord 1992). Referring to Anselin (1995), Getis-

Ord G_i^* is used in local spatial association indicators to identify local spatial clusters, where data points with spatial autocorrelation will be entered at neighboring points. There are two possibilities for spatial data patterns to become spatial autocorrelation issues; data with high values are close together (hot spots), and data with low values are close together (cold spots) (Ord and Getis 1995). The Getis-Ord G_i^* method is widely used in species distribution modeling and provides solutions to reduce spatial autocorrelation by assessing data distribution patterns (Cleasby et al. 2020; Cursach et al. 2020; Naimi et al. 2014; Pinault and Hunter 2011; Zhang et al. 2016). The hot spot analysis results produce 127 points free from spatial autocorrelation to reduce bias on the maxent model. Furthermore, environmental variables will be analyzed for multicollinearity first to separate if their environmental variables correlate (Baek et al. 2019; Li and Ding 2016). The following are the results of the multicollinearity test of ten environmental variables with 127 data on komodo's prey presences points, both directly and indirectly (Table 2).

Table 2 shows the results of the multicollinearity test between ten environmental variables, with the result that there are environmental variables that are correlated with a value less than -0.5 (negative) and more than 0.5 (positive) (McCarthy et al. 2015). Thus, the environmental variables free from multicollinearity will be further analyzed for spatial distribution modeling of maxent: moisture index, slope, distance from the settlement, and distance from the savanna.

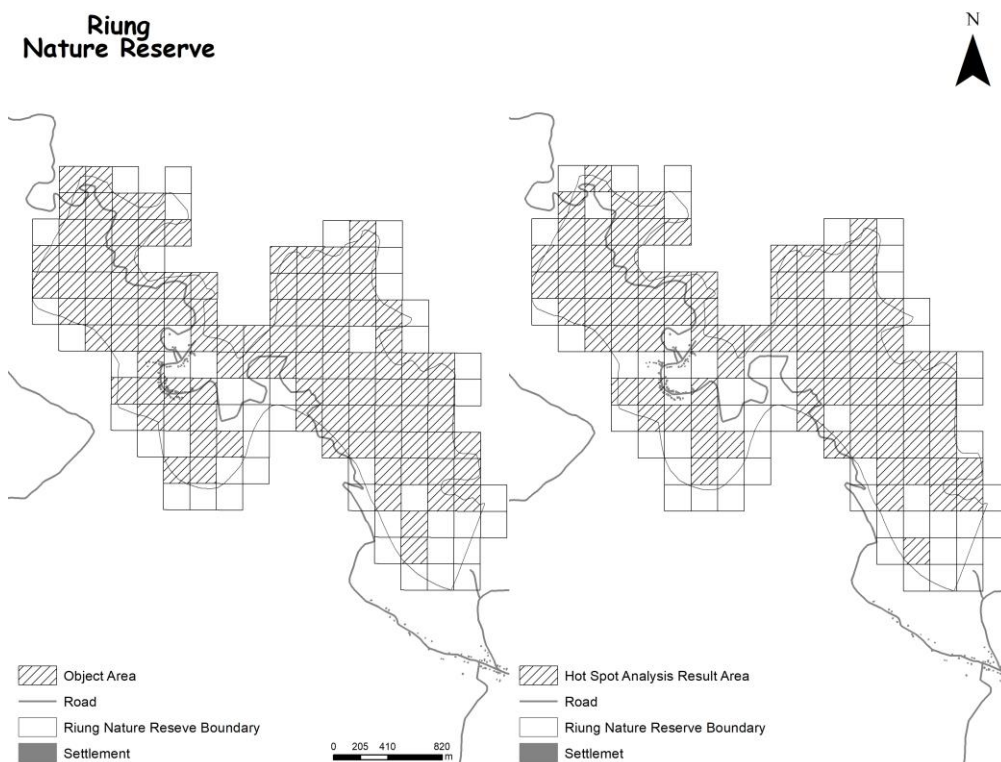


Figure 2. Distribution of Komodo dragon's potential prey in Riung based on survey object points and hot spot analysis result points

Maximum entropy result for Komodo dragon's potential prey

Moreover, with the presence of all Komodo prey, the modeling accuracy test based on the average sensitivity and specificity graphs (Figure 3) obtained an AUC (Area Under Curve) value of 0.776 with a standard deviation of 0.0414. Therefore, the AUC value indicates a modeling evaluation in the very good category.

The validation of the habitat suitability model produced by MaxEnt uses the AUC test value or the area under the ROC (Receiver Operating Curve) to identify the sensitivity and specificity of the model. In addition, the AUC value shows to observe the prediction accuracy of the distribution model (Lobo et al. 2008). Ultimately, the model will be accepted if the AUC value is >0.5 , indicating that the points of presence of species and variables overlap (Fielding and Bell 1997; Phillips 2005; Swets 1988).

Komodo dragon's potential prey in Riung is distributed in savanna, mangrove, and lowland forest. Based on Mustari et al. (2011), the Komodo dragon's prey are Timor deer, water buffalo, long-tailed macaque, wild horse, and wild boar in Rinca island. In Riung, six species are preyed on by Komodo dragons, Long-tailed macaque, house gecko, civet, Timor deer, horse, and cow. Cows and long-tailed macaques possibly became komodo prey and were found almost in Riung areas, while humans had settled horses and cows around the Riung nature reserve. In Riung, local people graze their cattle freely everywhere, so they are found in the nature reserve areas. That also allows komodo to prey on the cattle within their home-range area in the nature reserve of Riung. Still, we have no direct evidence that Komodo dragons prey on their cattle directly because of the difference between the activity of Komodo and monitoring time.

Table 2. The result of the multicollinearity test of ten environmental variables with 127 data points

Variable	EV1	EV2	EV3	EV4	EV5	EV6	EV7	EV8	EV9	EV10
EV1	1									
EV2	0.11624	1								
EV3	0.10242	0.05962	1							
EV4	-0.57065	0.01145	-0.30009	1						
EV5	-0.60820	-0.04200	-0.28469	0.92833	1					
EV6	0.01141	-0.35436	-0.03340	-0.12223	-0.02062	1				
EV7	0.61909	0.34173	0.09518	-0.48067	-0.51776	-0.17927	1			
EV8	0.71492	0.03176	0.27327	-0.74467	-0.77365	0.15562	0.55021	1		
EV9	-0.02709	-0.35529	-0.00438	-0.11878	-0.01341	0.98707	-0.22165	0.13362	1	
EV10	-0.11144	-0.47108	-0.07589	-0.07127	0.04819	0.63282	-0.17231	-0.00030	0.65845	1

Note: EV1 = elevation; EV2 = distance from savanna; EV3 = slope; EV4 = vegetation index; EV5 = moisture index; EV6 = distance from settlement; EV7 = distance from river; EV8 = land surface temperature; EV9 = distance from agriculture; EV10 = distance from road

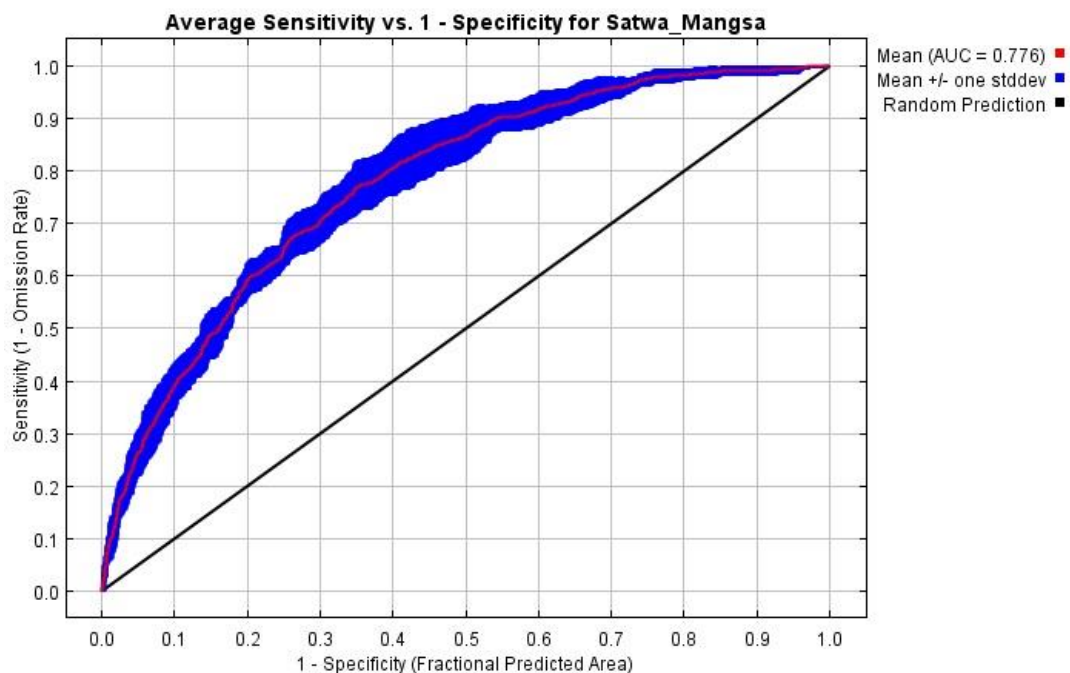


Figure 3. The AUC result of Komodo dragon's prey

Besides it, the presence of the Komodo dragon's potential prey was affected by environmental variables. However, based on the results of the jackknife AUC test (Figure 4), it is known that the environmental variable that has the most influence on performance in assessing modeling predictions is the moisture index. The range of values for the moisture index variable is from -1 - +1, where the greater can be categorized as having a high moisture content or a wet area, while the smaller value describes a dry area (Hardisky 1983; Xu 2005; Prasetyo 2017; Orimoloye 2020). Based on the moisture index, the data has a value range of -0.22 – +0.4 with a standard deviation of -0.054, found in savanna, lowland forest and mangrove habitats. At the study site, the moisture index range < 0 indicates a savanna habitat mixed with rocks, while the moisture index value > 0 indicates a savanna habitat dominated by grass and shrubs, lowland forest, and mangrove habitat. This is also explained in Xu (2005),

which states that a positive value on the moisture index describes a vegetated area and a negative value describes bare soil and buildings. The magnitude of the influence of environmental variables on modeling is also shown by the table of percent contribution (Table 3), which produces a habitat suitability map of Komodo dragon's potential prey in Figure 5.

Table 3. Percent contribution of each environmental variable to create a spatial distribution model

Variable	Percent contribution (%)
Moisture index	49.00
Distance from settlement	20.60
Distance from savanna	18.90
Slope	11.40

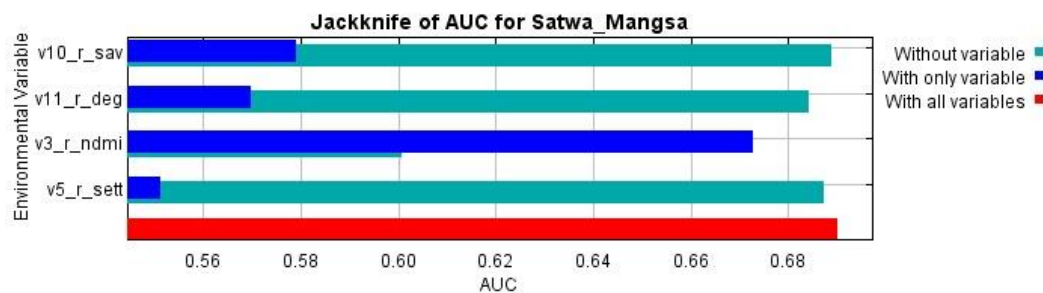


Figure 4. The AUC Jackknife for Komodo's prey

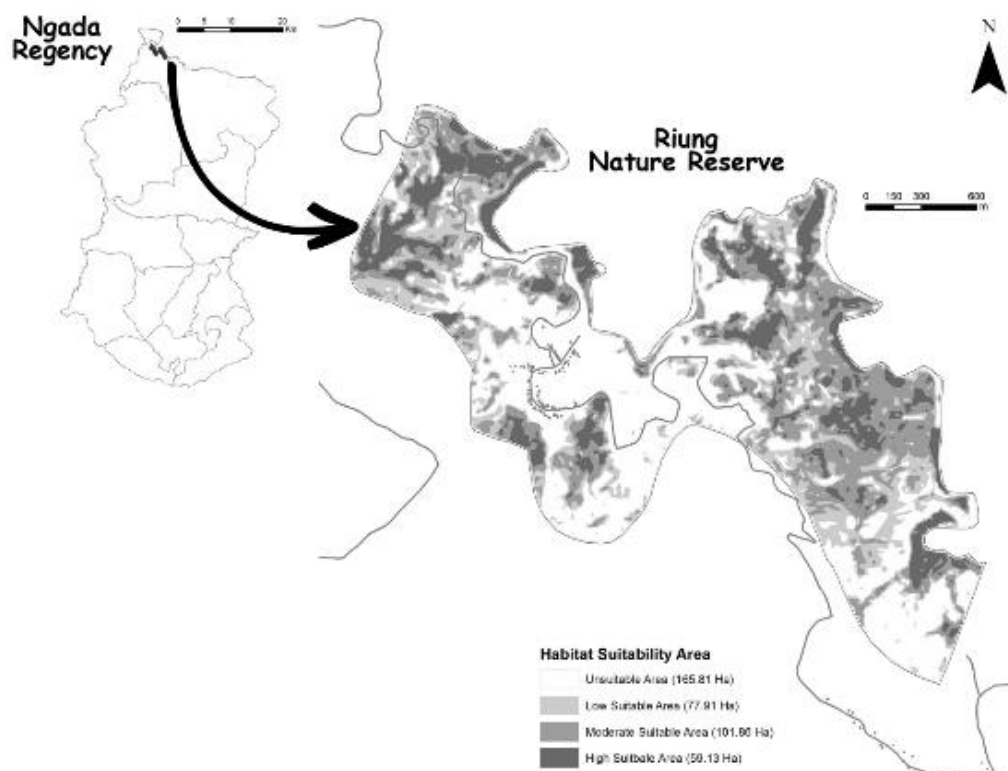


Figure 5. Habitat suitability map of Komodo dragon's potential prey

The results of Jackknife will show the contribution of each environmental variable to the model, indicating how important an environmental variable is to the existence of the Komodo dragon's potential prey. The final output from the results of spatial modeling of habitat suitability for Komodo dragon's potential prey is a habitat suitability map. The best chance of attendance is at the Habitat Suitability Index (HSI) value is close to 1, and the Maxent results are close to 0, indicating a low level of closeness/suitability.

The habitat suitability map shows areas with low (77.91 ha), moderate (101.86 ha), and high suitability (59.13 ha) for Komodo dragon's potential prey. In addition, there are areas with low suitability categories than high suitability. Therefore, low suitability indicates low habitat variables for carrying animals' needs.

Discussion

Habitat suitability links environmental factors and Komodo dragons to support the survival of animals. Guisan and Zimmermann (2000) reported that indirect variables are represented with spatial analysis in habitat areas. Therefore, a model combines abiotic, biotic, and habitat factors to build the link between environmental factors and animals. The variables for Komodo dragons selected were water distance, distance to human activities, and distance to the main road for representing the presence of the prey around human activities. Mostly in Riung, the prey and human overlapped for using resources. So, it is also possible for cattle to become prey to Komodo dragons, i.e., cows. In Riung, cattle were released anywhere that the cattle freely entered the conservation area. Therefore, the

overlapped area between wildlife animals such as Timor deer and cattle like cows is possible. In addition, the cattle preferred to choose a sloping area until moderate than a steep area. Khadka and James (2016) reported that slope and canopy areas affect animals' grazing choices. This result is also supported by Kayat et al. (2017) that there are different preferences between cattle and wildlife animals in the field. But Baskaran et al. (2016) reported different results from *Antilope cervicapra* Linnaeus, 1758 and *Equus caballus* Linnaeus, 1758, concentrated in similar grassland areas, and show the two species use together 11 out of 15 types of feed.

Prey's presence could affect Komodo dragons' behavior, actually movement behavior. That is linear with Sims et al. (2007, 2012) reported reptiles' movement to optimize prey foraging strategies. Purwandana et al. (2016) classified Komodo dragons into small dragons (<5 kg), medium dragons (15-25 kg), and large dragons (>50 kg) to understand that body mass affects prey preferences. Small Komodo dragons usually will consume small prey. Here are the differences between Komodo dragons in Riung and the main island of the komodo national park, which is about the body mass. Komodo dragons in Riung are smaller than Komodo dragons in Komodo Island. The potential prey for small Komodo dragons must include house geckos because geckos in the area overlap with Komodo dragons' habitat in Riung. Still, there is insufficient evidence showing Komodo dragons preying on geckos directly. We only found feces from the Komodo dragon's potential prey in all locations in Riung, showing that they can prey on house geckos (Figure 6).



Figure 6. Potential prey for Komodo dragon in Riung based on prey feces: A. Timor deer; B. Wild boar; C. Long-tailed macaque; D. Gecko

Other Komodo dragons' prey in the location is long-tailed macaques, civets, Timor deer, cows, and feral horses. That follows the results by Ariefiandy et al. (2013), which found three ungulates prey animals for the Komodo dragons: Timor deer, feral pig, and water buffalo. Previously, Mustari et al. (2011) reported that the potential prey for Komodo dragons are Timor deer, water buffalo, long-tailed macaque, wild horses, and wild boar. Among these five types of potential prey, three main prey are Komodo dragons, wild boars, Timor deer, and water buffalo (Ariefiandy 2011; Ariefiandy et al. 2013; Jessop et al. 2006). This is because the Komodo dragons selectively choose their prey according to the behavior of their prey. Long-tailed macaques, for example, are classified as difficult to catch because they move in groups and quickly, so they have good coordination to evade predators (Mustari et al. 2011). This statement is supported by Satria (2013), who states that Long-tailed macaques in the dry tropical forest of Rinca Island have a size of 40 individuals per group. Furthermore, Satria (2013) also reported the feeding habit of long-tailed macaques in all three age classes (adult males, adult females, and young) horizontally utilizing

more of the outermost tree canopy while vertically making more use of the forest floor.

The distribution pattern of the prey animals for the Komodo dragons follows the habitat type closely related to the availability of food for these prey animals. Timor deer and water buffalo are herbivores, so the presence of forage as feed is important in the existence and sustainability of these animal populations (Santosa et al. 2008; Masyud et al. 2007; Mustari et al. 2011). In addition to prey animals, residents' livestock, such as cows and goats, can run wild in the conservation area. These cattle in Riung exist because of human settlement (Figure 7), which could decrease the habitat's quality as a source of feed for mammals in wildlife (Sawadogo et al. 2005; Yoshihara et al. 2008; Baskaran et al. 2016). It also grazes in the habitat of the Komodo dragons, which could cause a decrease in the population of mammals as prey animals due to competition for resources with cattle (Mustari et al. 2010; zu Dohna et al. 2014). It is also a concern that this livestock may become dragons' prey even though they have not yet been found direct evidence that the Komodo dragons preyed on the cows and feral horses during the study.

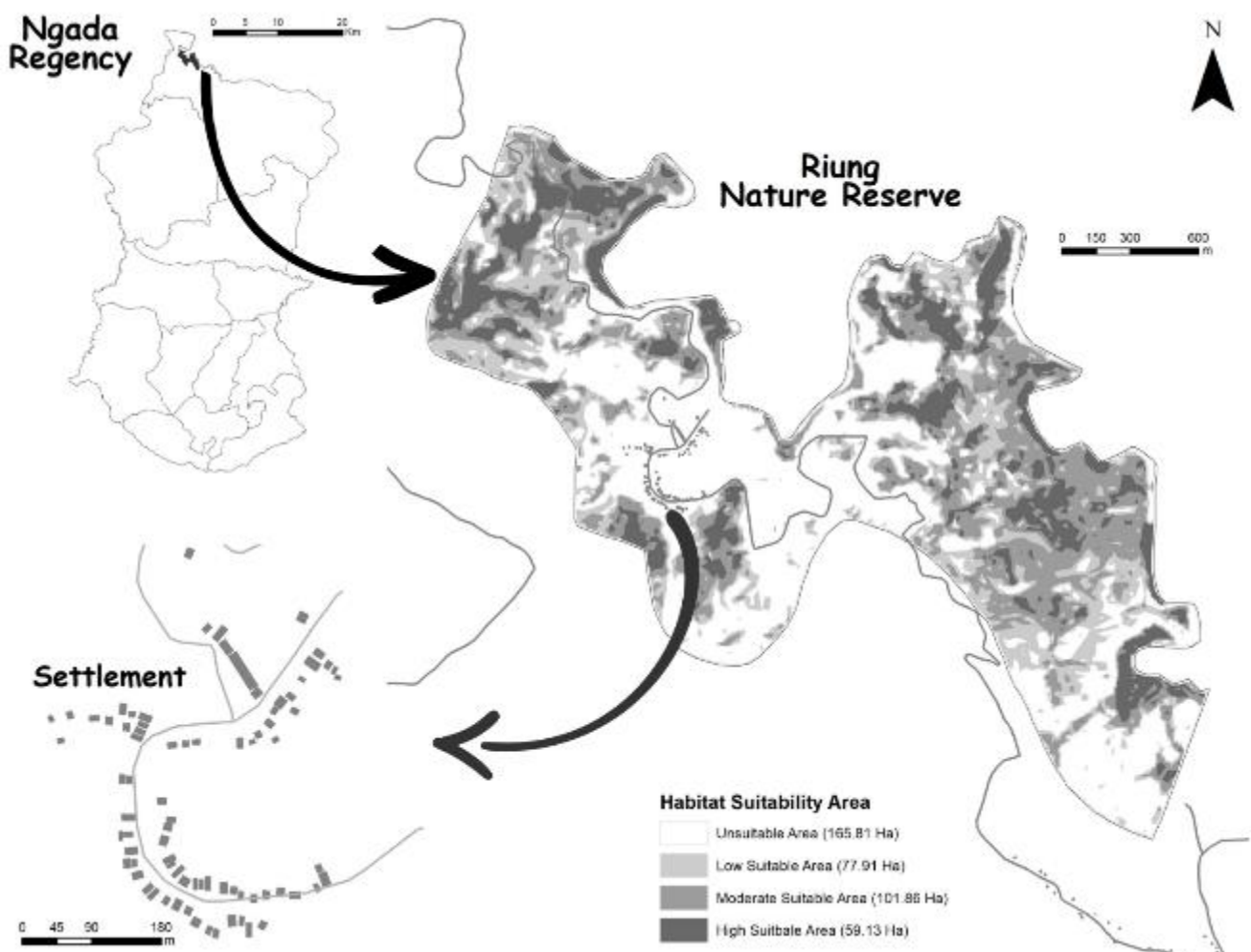


Figure 7. Human settlement in Riung Nature Reserve is close to moderate and low suitability areas of prey suitability habitat

The prey distribution location in Riung was also dominated by cattle presence; cows and horses were found in almost all locations in Riung. The cattle and their activities overlapped with the Komodo dragons' habitat and prey distribution prediction (Ariefiandy et al. 2021). Also, another threat to the presence of Komodo dragon is the conflict in this area is getting serious between the local community and the stakeholder. But on the other hand, we still find Komodo dragons in Riung around steep savanna with shrubs within the conflict area, but this discovery was not documented due to limited methods, tools, and time constraints.

We directly found a juvenile Komodo dragon around the hill close to the mangrove area but in the adjacent savanna; this area is hilly and dry but close to mangroves. Around this place, there is no water source for herbivores or ungulates. This Juvenile differs from Komodo on the main island of Komodo National Park. The Komodo has a lighter color, and the body is smaller than in the Komodo National Park. Therefore, the body size differences will also affect their prey size. For example, komodo dragons with body mass >18 kg will consume adult ungulates (Purwandana et al. 2021).

Wildlife animals, including Komodo dragons and their prey, tend not to harm humans and disturb human activities. In Riung, the local community lived in a conservation area close to low and moderate habitats suitable for prey. That may become a serious problem if conflicts between Komodo and the local community occur, possibly when the cattle become more potential prey to Komodo than another animal. This hypothesis can be true because, in Riung, most prey points' presences come from cattle than other ungulates. On the other hand, Komodo will not disturb human settlement unless there is no potential prey, such as cattle. The issue about human settlement and human activities in the Riung Nature Reserve area also affects the distribution of natural prey of Komodo dragons, such as Timor deer, caused by foraging competition in the grazing areas.

These results cannot explain whether a site suitable for prey animals is ideal for Komodo dragons. Referring to Ariefiandy et al. (2021), predicted distributions of Komodo dragons with the single-season occupancy models using the package unmarked intersect with Komodo dragons' prey suitable area in this study. It will be interesting to compare using the maximum entropy modeling with more detailed environmental variables according to Komodo dragons' ecology. Shadloo et al. (2021) reported in predicting habitat suitability for the desert monitor (*Varanus griseus* subsp *caspicus* Eichwald, 1831) using the Generalized Boosting Model (GBM), Generalized Additive Model (GAM), and Random Forest (RF) methods using ten environmental variables with a significant variable result is precipitation seasonality. Hosseinzadeh et al. (2020) also reported on the predicting past, current, and future habitat suitability and geographic distribution of the Iranian endemic species *Microgecko latifi* Leviton & Anderson, 1972 (Sauria: Geckonidae) using maxent analysis with 19 environmental variables and five variables selected after the correlation test produced a significant variable is precipitation of the

warmest quarter. Both studied the habitat suitability of lizard species with a significant bioclimatic variable. Further studies can be carried out to analyze the habitat suitability of lizards (*V. komodoensis*) by referring to (Hosseinzadeh et al. 2020; Shadloo et al. 2021) using physical and bioclimatic variables, and the results of Komodo dragons' potential prey habitat suitability studies can also be added as environmental variables with categorical data types.

ACKNOWLEDGEMENTS

We would like to thank Archipelagic Drylands Laboratory for financial research support and also thank the Natural Resources Conservation Agency of East Nusa Tenggara for the permit and support for us conducting this research through an assignment letter (ST. 680 /K.5.BIDTEK/PEG.3/09/2022). We thank and appreciate Dimas Christyan Pratama Arka, Ira Usman, David Daing, Rizal Floresto Irawan, and Manasye M. Manu throughout our research at Riung Nature Reserve Area.

REFERENCES

- Ab Lah NZ, Yusop Z, Hashim M, Mohd Salim J, Numata S. 2021. Predicting the habitat suitability of *Melaleuca cajuputi* based on the maxent species distribution model. *Forests* 12 (11): 1449. DOI: 10.3390/f12111449.
- Anenkhonov OA. 2009. Changes in the cenoflora of dark-coniferous forests of the Northern Baikal Region under current climate warming. *Geogr Nat Resour* 30 (4): 355-58. DOI: 10.1016/j.gnr.2009.11.009.
- Anselin L. 1995. Local indicators of spatial association-LISA. *Geogr Anal* 27 (2): 93-115. DOI: 10.1111/j.1538-4632.1995.tb00338.x.
- Ardiantiono, Jessop TS, Purwandana D, Ciofi C, Imansyah MJ, Panggur MR, Ariefiandy A. 2018. Effects of human activities on Komodo dragons in Komodo National Park. *Biodivers Conserv* 27: 3329-3347. DOI: 10.1007/s10531-018-1601-3.
- Ariefiandy A, Purwandana D, Azmi M, Nasu SA, Mardani J, Ciofi C, Jessop TS. 2021. Human activities associated with reduced Komodo dragon habitat use and range loss on Flores. *Biodivers Conserv* 30 (2): 461-479. DOI: 10.1007/s10531-020-02100-8.
- Ariefiandy A, Purwandana D, Coulson G, Forsyth DM, Jessop TS. 2013. Monitoring the ungulate prey of the Komodo dragon *Varanus komodoensis*: distance sampling or faecal counts? *Wildl Biol* 19 (2): 126-137. DOI: 10.2981/11-098.
- Ariefiandy A, Purwandana D, Nasu SA, Benu YJ, Chismiawati M, Kamil PI, Imansyah MJ, Ciofi C, Jessop T. 2017. Komodo Dragons Field Guide. Yayasan Komodo Survival Program, Denpasar. [Indonesian]
- Ariefiandy A. 2011. Population Assessments of Ungulate Prey and Komodo Dragons Across Protected Areas in Eastern Indonesia. [M.Phil. Thesis]. The University of Melbourne, Melbourne, Australia.
- Baek S, Kim MJ, Lee JH. 2019. Current and future distribution of *Ricania shantungensis* (Hemiptera: Ricaniidae) in Korea: Application of spatial analysis to select relevant environmental variables for MaxEnt and CLIMEX modeling. *Forests* 10 (6): 490. DOI: 10.3390/f10060490.
- Baker DJ, Maclean IMD, Goodall M, Gaston KJ. 2021. Species distribution modelling is needed to support ecological impact assessments. *J Appl Ecol* 58 (1): 21-26. DOI: 10.1111/1365-2664.13782.
- Baral S, Kunwar A, Adhikari D, Kandel K, Mandal DN, Thapa A, Neupane D, Thapa TB. 2023. The potential distribution of the yellow monitor, *Varanus flavescens* (Hardwick & Gray) under multiple climate, land cover and dispersal scenarios in Nepal. *CSIRO, Wildl Res*, Australia. DOI: 10.1071/WR22176.
- Baskaran N, Ramkumaran K, Karthikeyan G. 2016. Spatial and dietary overlap between blackbuck (*Antelope cervicapra*) and feral horse

- (*Equus caballus*) at point calimere wildlife sanctuary, Southern India: Competition between native versus introduced species. *Mammal Biol* 81: 295-302. DOI: 10.1016/j.mambio.2016.02.004.
- Bellard C, Bertelsmeier C, Leadley P, Thuiller W, Courchamp F. 2012. Impacts of climate change on the future of biodiversity. *Ecol Lett* 15 (4): 365-377. DOI: 10.1111/j.1461-0248.2011.01736.x.
- Bestion E, Teyssier A, Richard M, Clobert J, Cote J. 2015. Live fast, die young: experimental evidence of population extinction risk due to climate change. *PLoS Biol* 13 (10): e1002281. DOI: 10.1371/journal.pbio.1002281.
- Bismark M. 2011. Standard Operating Procedures for species diversity surveys in protected areas. Balitbang Kehutanan, Bogor. [Indonesian]
- Brown JL, Bennett JR, French CM. 2017. SDMtoolbox 2.0: the next generation Python-based GIS toolkit for landscape genetic, biogeographic and species distribution model analyses. *PeerJ* 5: e4095. DOI: 10.7717/peerj.4095.
- Brown JL. 2014. SDM toolbox: a python-based GIS toolkit for landscape genetic, biogeographic and species distribution model analyses. *Methods Ecol Evol* 5 (7): 694-700. DOI: 10.1111/2041-210X.12200.
- Cabrera JS, Lee HS. 2020. Flood risk assessment for Davao Oriental in the Philippines using geographic information system-based multi-criteria analysis and the maximum entropy model. *J Flood Risk Manag* 13 (2): e12607. DOI: 10.1111/jfr3.12607.
- Che L, Cao B, Bai C, Wang J, Zhang L. 2014. Predictive distribution and habitat suitability assessment of *Notholirion bulbuliferum* based on MaxEnt and ArcGIS. *Chin J Ecol* 33 (6): 1623.
- Chefaoui RM, Hosseinzadeh MS, Mashayekhi M, Safaei-Mahroo B, Kazemi SM. 2018. Identifying suitable habitats and current conservation status of a rare and elusive reptile in Iran. *Amphibia-Reptilia* 39 (3): 355-362. DOI: 10.1163/15685381-17000185.
- Chen IC, Hill JK, Ohlemüller R, Roy DB, Thomas CD. 2011. Rapid range shifts of species associated with high levels of climate warming. *Science* 333 (6045): 1024-1026. DOI: 10.1126/science.1206432.
- Ciofi C, De Boer ME. 2004. Distribution and conservation of the Komodo monitor (*Varanus komodoensis*). *Herpetol J* 14 (2): 99-107.
- Cleasby IR, Owen E, Wilson L, Wakefield ED, O'Connell P, Bolton M. 2020. Identifying important at-sea areas for seabirds using species distribution models and hotspot mapping. *Biol Conserv* 241: 108375. DOI: 10.1016/j.biocon.2019.108375.
- Çoban HO, Örcüçü ÖK, Arslan ES. 2020. MaxEnt modeling for predicting the current and future potential geographical distribution of *Quercus libani* Olivier. *Sustainability* 12 (7): 2671. DOI: 10.3390/su12072671.
- Cuena-Lombrana A, Fois M, Fenu G, Cogoni D, Bacchetta G. 2018. The impact of climatic variations on the reproductive success of *Gentiana lutea* L. in a Mediterranean mountain area. *Int J Intl Biometeorol* 62: 1283-1295. DOI: 10.1007/s00484-018-1533-3.
- Cursach J, Far AJ, Ruiz M. 2020. Geospatial analysis to assess distribution patterns and predictive models for endangered plant species to support management decisions: a case study in the Balearic Islands. *Biodivers Cons* 29: 3393-3410. DOI: 10.1007/s10531-020-02029-y.
- Deniau C, Gaillard T, Mbagogo A, Réounodji F, Le Bel S. 2017. Using the KoBoCollect tool to analyze the socio-economic and socio-cultural aspects of commercial hunting and consumption of migratory waterbirds in the Lakes Chad and Fitri. Conference Proceedings of 2017 EFITA WCCA congress: European conference dedicated to the future use of ICT in the agri-food sector, bioresource and biomass sector. Irstea, Montpellier, 2-6 July 2017.
- Dillon ME, Wang G, Huey RB. 2010. Global metabolic impacts of recent climate warming. *Nature* 467 (7316): 704-706. DOI: 10.1038/nature09407.
- Ehrlén J, Morris WF. 2015. Predicting changes in the distribution and abundance of species under environmental change. *Ecol Lett* 18 (3): 303-314. DOI: 10.1111/ele.12410.
- Elith J, Graham CH, Anderson RP, Dudík M, Ferrier S, Guisan A, Hijmans RJ, Huettmann F, Leathwick JR, Lehmann A, Li J, Lohmann LG, Loiselle BA, Manion G, Moritz C, Nakamura M, Nakazawa Y, Overton JMcM, Peterson AT, Phillips SJ, Richardson K, Pereira RS, Schapire RE, Soberón J, Williams S, Wisz MS, Zimmermann NE. 2006. Novel methods improve prediction of species' distributions from occurrence data. *Ecography* 29 (2): 129-151. DOI: 10.1111/j.2006.0906-7590.04596.x.
- Elith J, Kearney M, Phillips S. 2010. The art of modelling range-shifting species. *Methods Ecol Evol* 1 (4): 330-342. DOI: 10.1111/j.2041-210X.2010.00036.x.
- Elith J, Phillips SJ, Hastie T, Dudík M, Chee YE, Yates CJ. 2011. A statistical explanation of MaxEnt for ecologists. *Divers Distrib* 17 (1): 43-57. DOI: 10.1111/j.1472-4642.2010.00725.x.
- Fathinia B, Rödder D, Rastegar-Pouyani N, Rastegar-Pouyani E, Hosseinzadeh MS, Kazem, SM. 2020. The past, current and future habitat range of the Spider-tailed Viper, *Pseudocerastes urarachnoides* (Serpentes: Viperidae) in western Iran and eastern Iraq as revealed by habitat modelling. *Zool Middle East* 66 (3): 197-205. DOI: 10.1080/09397140.2020.1757910.
- Fauzia AM. 2020. Komodo Dragon Behavior in the Tourism and Non-tourism Areas of Rinca Island, Komodo National Park, East Nusa Tenggara. [Undergraduate Thesis]. Institut Pertanian Bogor, Bogor. [Indonesian]
- Fielding AH, Bell JF. 1997. A review of methods for the assessment of prediction errors in conservation presence/absence models. *Environ Conserv* 24 (1): 38-49. DOI: 10.1017/S0376892997000088.
- Fois M, Cuena-Lombrana A, Fenu G, Cogoni D, Bacchetta G. 2018. Does a correlation exist between environmental suitability models and plant population parameters? An experimental approach to measure the influence of disturbances and environmental changes. *Ecol Indic* 86: 1-8. DOI: 10.1016/j.ecolind.2017.12.009.
- Franklin J. 2010. Mapping Species Distributions: Spatial Inference and Prediction. Cambridge University Press, Cambridge. DOI: 10.1017/CBO9780511810602.
- Getis A, Ord JK. 1992. The analysis of spatial association by use of distance statistics. *Geogr Anal* 24 (3): 189-206. DOI: 10.1111/j.1538-4632.1992.tb00261.x.
- Graham CH, Ferrier S, Huettman F, Moritz C, Peterson AT. 2004. New developments in museum-based informatics and applications in biodiversity analysis. *Trends Ecol Evol* 19 (9): 497-503. DOI: 10.1016/j.tree.2004.07.006.
- Guisan A, Thuiller W. 2005. Predicting species distribution: offering more than simple habitat models. *Ecol Lett* 8 (9): 993-1009. DOI: 10.1111/j.1461-0248.2005.00792.x.
- Guisan A, Zimmermann NE. 2000. Predictive habitat distribution models in ecology. *Ecol Model* 135 (2-3): 147-186. DOI: 10.1016/S0304-3800(00)00354-9.
- Habibzadeh N, Ghoddousi A, Bleyhl B, Kuemmerle T. 2021. Rear-edge populations are important for understanding climate change risk and adaptation potential of threatened species. *Conserv Sci Pract* 3 (5): e375. DOI: 10.1111/csp2.375.
- Hardisky MA, Klemas V, Smart M. 1983. The influence of soil salinity, growth form, and leaf moisture on the spectral radiance of *Spartina alterniflora*. *Photogramm Eng Remote Sens* 49: 77-83.
- Hosseinzadeh MS, Farhadi QM, Naimi B, Roedder D, Kazemi SM. 2018. Habitat suitability and modelling the potential distribution of the Plateau Snake Skink *Ophiomorus nuchalis* (Sauria Scincidae) on the Iranian Plateau. *North-Western J Zool* 14 (1): 60-63.
- Hosseinzadeh MS, Fois M, Zangi B, Kazemi SM. 2020. Predicting past, current and future habitat suitability and geographic distribution of the Iranian endemic species *Microgecko latifi* (Sauria: Gekkonidae). *J Arid Environ* 183: 104283. DOI: 10.1016/j.jaridenv.2020.104283.
- Hosseinzadeh MS, Ghezellou P, Kazemi SM. 2017. Predicting the potential distribution of the endemic snake *Spalerosophis microlepis* (Serpentes: Colubridae), in the Zagros Mountains, western Iran. *Salamandra* 53 (2): 294-298.
- Iverson LR, McKenzie D. 2013. Tree-species range shifts in a changing climate: detecting, modeling, assisting. *Landsc Ecol* 28: 879-889. DOI: 10.1007/s10980-013-9885-x.
- Jessop T, Ariefiandy A, Azmi M, Ciofi C, Imansyah J, Purwandana D. 2021. *Varanus komodoensis*. The IUCN Red List of Threatened Species 2021: e.T22884A123633058. DOI: 10.2305/IUCN.UK.2021-2.RLTS.T22884A123633058.en.
- Jessop TS, Madsen T, Ciofi C, Imansyah MJ, Purwandana D, Rudiharto H, Ariefiandy A, Phillips JA. 2007. Island differences in population size structure and catch per unit effort and their conservation implications for Komodo dragons. *Biol Conserv* 135 (2): 247-255. DOI: 10.1016/j.biocon.2006.10.025.
- Jessop TS, Madsen T, Sumner J, Rudiharto H, Phillips JA, Ciofi C. 2006. Maximum body size among insular Komodo dragon populations covaries with large prey density. *Oikos* 112 (2): 422-429. DOI: 10.1111/j.0030-1299.2006.14371.x.
- Kaky E, Gilbert F. 2017. Predicting the distributions of Egypt's medicinal plants and their potential shifts under future climate change. *PLoS One* 12 (11): e0187714. DOI: 10.1371/journal.pone.0187714.

- Kaky E, Gilbert F. 2019. Assessment of the extinction risks of medicinal plants in Egypt under climate change by integrating species distribution models and IUCN Red List criteria. *J Arid Environ* 170: 103988. DOI: 10.1016/j.jaridenv.2019.05.016.
- Kayat K, Pudyatmoko S, Maksun M, Imron, MA. 2017. Potential conflict of horse herding in the habitat of the Timor deer (*Rusa timorensis* Blainville 1822) in Tanjung Torong Padang area, East Nusa Tenggara. *Jurnal Ilmu Kehutanan* 11 (1): 4-18. DOI: 10.22146/jik.24866. [Indonesian]
- Keith DA, Akçakaya HR, Thuiller W, Midgley GF, Pearson RG, Phillips SJ, Regan HM, Araújo MB, Rebelo TG. 2008. Predicting extinction risks under climate change: coupling stochastic population models with dynamic bioclimatic habitat models. *Biol Lett* 4 (5): 560-563. DOI: 10.1098/rsbl.2008.0049.
- Kennedy UF, Kusri MD, Ariefiandy A, Mardiatuti A. 2020. Invasive toads are close to but absent from Komodo National Park. *BIO Web Conf* 19: 00017. DOI: 10.1051/bioconf/20201900017.
- Khadka KK, James DA. 2016. Habitat selection by endangered Himalayan musk deer (*Moschus chrysogaster*) and impacts of livestock grazing in Nepal Himalaya: Implications for conservation. *J Nat Conserv* 31: 38-42. DOI: 10.1016/j.jnc.2016.03.002.
- Khanum R, Mumtaz AS, Kumar S. 2013. Predicting impacts of climate change on medicinal asclepiads of Pakistan using Maxent modeling. *Acta Oecolog* 49: 23-31. DOI: 10.1016/j.actao.2013.02.007.
- Kornejady A, Ownegh M, Bahremand A. 2017. Landslide susceptibility assessment using maximum entropy model with two different data sampling methods. *Catena* 152: 144-162. DOI: 10.1016/j.catena.2017.01.010.
- Lakshminarasimhappa MC. 2022. Web-based and smart mobile app for data collection: Kobo Toolbox/Kobo collect. *J Indian Libr Assoc* 57 (2): 72-79.
- Le Bel S, Chavernac D, Stansfield F. 2016. Promoting a mobile data collection system to improve HWC incident recording: a simple and handy solution for controlling problem animals in southern Africa. In: Angelici F (eds). *Problematic Wildlife*. Springer, Cham. DOI: 10.1007/978-3-319-22246-2_19.
- Li S, Wang Z, Zhu Z, Tao Y, Xiang J. 2023. Predicting the potential suitable distribution area of *Emeia pseudosauteri* in Zhejiang Province based on the MaxEnt model. *Sci Rep* 13 (1): 1806. DOI: 10.1038/s41598-023-29009-w.
- Li Y, Ding C. 2016. Effects of sample size, sample accuracy and environmental variables on predictive performance of MaxEnt model. *Polish J Ecol* 64 (3): 303-312. DOI: 10.3161/15052249PJE2016.64.3.001.
- Lobo JM, Jiménez-Valverde A, Real R. 2008. AUC: a misleading measure of the performance of predictive distribution models. *Glob Ecol Biogeogr* 17 (2): 145-151. DOI: 10.1111/j.1466-8238.2007.00358.x.
- Ma B, Sun J. 2018. Predicting the distribution of *Stipa purpurea* across the Tibetan Plateau via the MaxEnt model. *BMC Ecol* 18: 1-12. DOI: 10.1186/s12898-018-0165-0.
- Malakhov DV, Chirikova MA. 2018. Species distribution model of *Varanus griseus caspius* (Eichwald, 1831) in Central Asia: An insight to the species'biology. *Russian J Herpetol* 25 (3): 195-206. DOI: 10.30906/1026-2296-2018-25-3-195-206.
- Masyud B, Wijaya R, Santos, IB. 2007. Distribution, population and daily activities of Timor deer (*Cervus timorensis*, de Blainville 1822) in Bali Barat National Park. *Media Konservasi* 12 (3): DOI: 10.29244/medkon.12.3.%25p. [Indonesian]
- McCarthy JL, Wibisono HT, McCarthy KP, Fuller TK, Andayani N. 2015. Assessing the distribution and habitat use of four felid species in Bukit Barisan Selatan National Park, Sumatra, Indonesia. *Glob Ecol Conserv* 3: 210-221. DOI: 10.1016/j.gecco.2014.11.009.
- Merilä J, Hendry AP. 2014. Climate change, adaptation, and phenotypic plasticity: the problem and the evidence. *Evol Appl* 7 (1): 1-14. DOI: 10.1111/eva.12137.
- Merow C, Smith MJ, Silander Jr JA. 2013. A practical guide to MaxEnt for modeling species' distributions: what it does, and why inputs and settings matter. *Ecography* 36 (10): 1058-1069. DOI: 10.1111/j.1600-0587.2013.07872.x.
- Mustari AH, Djuanda TD, Sihite J. 2011. The potential of large mammals as prey for Komodo dragons (*Varanus komodoensis* Ouwens 1912) in Rinca Island, Komodo National Park, East Nusa Tenggara. *Media Konservasi* 16 (1): 47-53. DOI: 10.29244/medkon.16.1.%25p. [Indonesian]
- Mustari AH, Siga HR, Noviandi T. 2010. Ecological study and existence status of Komodo dragons (*Varanus komodoensis*) in Padar Island, Komodo National Park. *Media Konservasi* 15 (1): 13-20. DOI: 10.29244/medkon.15.1.%25p. [Indonesian]
- Na X, Zhou H, Zang S, Wu C, Li W, Li M. 2018. Maximum entropy modeling for habitat suitability assessment of Red-crowned crane. *Ecol Indic* 91: 439-446. DOI: 10.1016/j.ecolind.2018.04.013.
- Naimi B, Hamm NAS, Groen TA, Skidmore AK, Toxopeus AG. 2014. Where is positional uncertainty a problem for species distribution modelling? *Ecography* 37 (2): 191-203. DOI: 10.1111/j.1600-0587.2013.00205.x.
- Nenzén HK, Araújo MB. 2011. Choice of threshold alters projections of species range shifts under climate change. *Ecol Model* 222 (18): 3346-3354. DOI: 10.1016/j.ecolmodel.2011.07.011.
- Newbold T. 2018. Future effects of climate and land-use change on terrestrial vertebrate community diversity under different scenarios. *Proc R Soc B* 285 (1881): 20180792. DOI: 10.1098/rspb.2018.0792.
- Nezer O, Bar-David S, Gueta T, Carmel Y. 2017. High-resolution species-distribution model based on systematic sampling and indirect observations. *Biodivers Conserv* 26: 421-437. DOI: 10.1007/s10531-016-1251-2.
- Ord JK, Getis A. 1995. Local spatial autocorrelation statistics: distributional issues and an application. *Geogr Anal* 27 (4): 286-306. DOI: 10.1111/j.1538-4632.1995.tb00912.x.
- Orimoloye IR., Oloade OO. 2020. Spatial evaluation of land-use dynamics in gold mining area using remote sensing and GIS technology. *Intl J Environ Sci Technol* 17: 4465-4480. DOI: 10.1007/s13762-020-02789-8.
- Oxoli D, Prestifilippo G, Bertocchi D. 2017. Enabling spatial autocorrelation mapping in QGIS: The hotspot analysis Plugin. *GEAM. Geoinformatica Ambientale E Mineraria* 151 (2): 45-50.
- Palla F, Le Bel S, Chavernac D, Cornélis D. 2016. New technologies: mobile data collection system implication for wildlife management in Central Africa. *Proceedings of Annual Meeting of the Association for Tropical Biology and Conservation Tropical Ecology and Society Reconciliating Conservation and Sustainable Use of Biodiversity*, Montpellier, 19-23 June 2016.
- Perkins-Taylor IE, Frey JK. 2020. Predicting the distribution of a rare chipmunk (*Neotamias quadrivittatus oscuraensis*): comparing MaxEnt and occupancy models. *J Mamm* 101 (4): 1035-1048. DOI: 10.1093/jmammal/gyaa057.
- Phillips SJ, Anderson RP, Dudík M, Schapire RE, Blair ME. 2017. Opening the black box: An open-source release of Maxent. *Ecography* 40 (7): 887-893. DOI: 10.1111/ecog.03049.
- Phillips SJ, Anderson RP, Schapire RE. 2006. Maximum entropy modeling of species geographic distributions. *Ecol Model* 190 (3-4): 231-259. DOI: 10.1016/j.ecolmodel.2005.03.026.
- Phillips SJ, Dudík M. 2008. Modeling of species distributions with Maxent: new extensions and a comprehensive evaluation. *Ecography* 31 (2): 161-175. DOI: 10.1111/j.0906-7590.2008.5203.x.
- Phillips SJ. 2005. A brief tutorial on Maxent. *AT&T Res* 190 (4): 231-259.
- Phillips SJ. 2017. A brief tutorial on Maxent. *AT&T Research*.
- Pinault LL, Hunter FF. 2011. New highland distribution records of multiple Anopheles species in the Ecuadorian Andes. *Malaria J* 10: 1-11. DOI: 10.1186/1475-2875-10-236.
- Prasetyo LB. 2017. Pendekatan Ekologi Lanskap untuk Konservasi Biodiversitas. *Fakultas Kehutanan Institut Pertanian Bogor, Bogor*. [Indonesian]
- Purwandana D, Ariefiandy A, Imansyah MJ, Rudiharto H, Seno A, Ciofi C, Fordham DA, Jessop TS. 2014. Demographic status of Komodo dragons populations in Komodo National Park. *Biol Conserv* 171: 29-35. DOI: 10.1016/j.biocon.2014.01.017.
- Purwandana D, Ariefiandy A, Imansyah MJ, Seno A, Ciofi C, Letnic M, Jessop TS. 2016. Ecological allometries and niche use dynamics across Komodo dragon ontogeny. *Sci Nat* 103: 1-11. DOI: 10.1007/s00114-016-1351-6.
- Purwandana D, Ciofi C, Imansyah MJ, Ariefiandy A, Rudiharto H, Jessop TS. 2021. Prey preferences and body mass most influence movement behavior and home range area of Komodo Dragons. *Ichtiol Herpetol* 109 (1): 92-101. DOI: 10.1643/h2020028.
- Rahmati O, Pourghasemi HR, Melesse AM. 2016. Application of GIS-based data driven random forest and maximum entropy models for groundwater potential mapping: a case study at Mehran Region, Iran. *Catena* 137: 360-372. DOI: 10.1016/j.catena.2015.10.010.
- Remya K, Ramachandran A, Jayakumar S. 2015. Predicting the current and future suitable habitat distribution of *Myristica dactyloides*

- Gaertn. using MaxEnt model in the Eastern Ghats, India. *Ecol Eng* 82: 184-188. DOI: 10.1016/j.ecoleng.2015.04.053.
- Santosa Y, Auliyani D, Kartono AP. 2008. Estimation of growth model and spatial distribution of Timor deer population (*Cervus timorensis* de Blainville, 1822) in Alas Purwo National Park, East Java. *Media Konservasi* 13 (1): 1-7. DOI: 10.29244/medkon.13.1.%25p. [Indonesian]
- Santosa Y, Muhammad RYZ, Rahman DA. 2012. Estimation of demographic parameters and spatial distribution of Komodo dragons on Rinca Island, Komodo National Park. *Jurnal Ilmu Pertanian Indonesia* 17 (2): 126-131. [Indonesian]
- Satria RA, Imron MA. 2013. Feeding Behavior of Long-Tailed Monkeys (*Macaca fascicularis*) in Dry Tropical Forest Ecosystems on Rinca Island, Komodo National Park. [Undergraduate Thesis]. Universitas Gadjah Mada, Yogyakarta. [Indonesian]
- Sawadogo L, Tiveau D, Nygård R. 2005. Influence of selective tree cutting, livestock and prescribed fire on herbaceous biomass in the savannah woodlands of Burkina Faso, West Africa. *Agric Ecosyst Environ* 105 (1-2): 335-345. DOI: 10.1016/j.agee.2004.02.004.
- Shadloo S, Mahmoodi S, Hosseinzadeh MS, Kazemi SM. 2021. Prediction of habitat suitability for the desert monitor (*Varanus griseus caspius*) under the influence of future climate change. *J Arid Environ* 186: 104416. DOI: 10.1016/j.jaridenv.2020.104416.
- Sharma S, Arunachalam K, Bhavsar D, Kala R. 2018. Modeling habitat suitability of *Perilla frutescens* with MaxEnt in Uttarakhand-A conservation approach. *J Appl Res Med Aromat Plant* 10: 99-105. DOI: 10.1016/j.jarmap.2018.02.003.
- Shrestha N. 2020. Detecting multicollinearity in regression analysis. *Am J Appl Math Stat* 8 (2): 39-42. DOI: 10.12691/ajams-8-2-1.
- Sims DW, Humphries NE, Bradford RW, Bruce, BD. 2012. Lévy flight and Brownian search patterns of a free-ranging predator reflect different prey field characteristics. *J Anim Ecol* 81 (2): 432-442. DOI: 10.1111/j.1365-2656.2011.01914.x.
- Sims DW, Righton D, Pitchford JW. 2007. Minimizing errors in identifying Lévy flight behaviour of organisms. *J Anim Ecol* 76 (2): 222-229. DOI: 10.1111/j.1365-2656.2006.01208.x.
- Songchitruksa P, Zeng X. 2010. Getis-Ord spatial statistics to identify hot spots by using incident management data. *Transp Res Record* 2165 (1): 42-51. DOI: 10.3141/2165-05.
- Su H, Bista M, Li M. 2021. Mapping habitat suitability for Asiatic black bear and red panda in Makalu Barun National Park of Nepal from Maxent and GARP models. *Sci Rep* 11 (1): 1-14. DOI: 10.1038/s41598-021-93540-x.
- Sun X, Long Z, Jia J. 2021. A multi-scale Maxent approach to model habitat suitability for the giant pandas in the Qionglai mountain, China. *Glob Ecol Conserv* 30: e01766. DOI: 10.1016/j.gecco.2021.e01766.
- Swets JA. 1988. Measuring the accuracy of diagnostic systems. *Science* 240 (4857): 1285-1293. DOI: 10.1126/science.3287615.
- Thomas CD, Cameron A, Green RE, Bakkenes M, Beaumont LJ, Collingham YC, Erasmus BFN, De Siqueira MF, Grainger A, Hannah L. 2004. Extinction risk from climate change. *Nature* 427 (6970): 145-148. DOI: 10.1038/nature02121.
- Ujvari B, Mun HC, Conigrave AD, Ciofi C, Madsen T. 2015. Invasive toxic prey may imperil the survival of an iconic giant lizard, the Komodo dragon. *Pac Conserv Biol* 20 (4): 363-365. DOI: 10.1071/PC140363.
- Walpole MJ. 2001. Feeding dragons in Komodo National Park: a tourism tool with conservation complications. *Anim Conserv* 4 (1): 67-73. DOI: 10.1017/S136794300100107X.
- Wang WJ, He HS, Thompson III FR, Spetich MA, Fraser JS. 2018. Effects of species biological traits and environmental heterogeneity on simulated tree species distribution shifts under climate change. *Sci Tot Environ* 634: 1214-1221. DOI: 10.1016/j.scitotenv.2018.03.353.
- Williams KJ, Belbin L, Austin MP, Stein JL, Ferrier S. 2012. Which environmental variables should I use in my biodiversity model? *Intl J Geogr Inf Sci* 26 (11): 2009-2047. DOI: 10.1080/13658816.2012.698015.
- Wisz MS, Hijmans RJ, Li J, Peterson AT, Graham CH, Guisan A. 2008. NCEAS predicting species distributions working group. *Divers Distrib* 14 (5): 763-773. DOI: 10.1111/j.1472-4642.2008.00482.x.
- Wolkovich EM, Cook BI, Allen JM, Crimmins TM, Betancourt JL, Travers SE, Pau S, Regetz J, Davies TJ, Kraft NJB. 2012. Warming experiments underpredict plant phenological responses to climate change. *Nature* 485 (7399): 494-497. DOI: 10.1038/nature11014.
- Xu H. 2005. A study on information extraction of water body with the modified normalized difference water index (MNDWI). *J Remote Sens* 9 (5): 595.
- Yang XQ, Kushwaha SPS, Saran S, Xu J, Roy PS. 2013. Maxent modeling for predicting the potential distribution of medicinal plant, *Justicia adhatoda* L. in Lesser Himalayan foothills. *Ecol Eng* 51: 83-87. DOI: 10.1016/j.ecoleng.2012.12.004.
- Yoshihara Y, Ito TY, Lhagvasuren B, Takatsuki S. 2008. A comparison of food resources used by Mongolian gazelles and sympatric livestock in three areas in Mongolia. *J Arid Environ* 72 (1): 48-55. DOI: 10.1016/j.jaridenv.2007.05.001.
- Yu W, Ai T, Yang M, Liu J. 2016. Detecting "Hot Spots" of facility POIs based on kernel density estimation and spatial autocorrelation technique. *武汉大学学报• 信息科学版* 41 (2): 221-227. DOI: 10.13203/j.whugis.20140092.
- Zhang D, Sun C, Yu H, Li J, Liu W, Li Z, Bao C, Liu D, Zhang N, Zhu F. 2019a. Environmental risk factors and geographic distribution of severe fever with thrombocytopenia syndrome in Jiangsu Province, China. *Vector Borne Zoonotic Dis* 19 (10): 758-766. DOI: 10.1089/vbz.2018.2425.
- Zhang J, Jiang F, Li G, Qin W, Li S, Gao H, Cai Z, Lin G, Zhang T. 2019b. Maxent modeling for predicting the spatial distribution of three raptors in the Sanjiangyuan National Park, China. *Ecol Evol* 9 (11): 6643-6654. DOI: 10.1002/ece3.5243.
- Zhang MG, Slik JW, Ma KP. 2016. Using species distribution modeling to delineate the botanical richness patterns and phytogeographical regions of China. *Sci Rep* 6 (1): 1-9. DOI: 10.1038/srep22400.
- zu Dohna H, Peck DE, Johnson BK, Reeves A, Schumaker BA. 2014. Wildlife-livestock interactions in a western rangeland setting: Quantifying disease-relevant contacts. *Prev Vet Med* 113 (4): 447-456. DOI: 10.1016/j.prevetmed.2013.12.004.